

Tableau de situation du Comité technique de l'ours blanc et renseignements à l'appui pour 2024

28 mai 2024

Contenu

TERMES UTILISÉS DANS LE TABLEAU DE SITUATION DU CTOB POUR 2024.....	2
TABLEAU DE SITUATION DU CTOB POUR 2024.....	9
TABLEAU DES PRÉLÈVEMENTS DU CTOB POUR 2024	14
DESCRIPTIONS DES SOUS-POPULATIONS D'OURS BLANC DU CTOB POUR 2024.....	15
BAIE DE BAFFIN (BB)	15
DÉTROIT DE DAVIS (DS)	17
BASSIN FOXE (BF)	20
GOLFE DE BOOTHIA (GB).....	22
BASSIN KANE (BK)	24
DÉTROIT DE LANCASTER (DL).....	25
DÉTROIT DE M'CLINTOCK (DM).....	27
NORD DE LA MER DE BEAUFORT (NMB).....	29
BAIE NORWEGIAN (BN)	33
SUD DE LA MER DE BEAUFORT (SMB)	34
SUD DE LA BAIE D'HUDSON (SBH).....	38
DÉTROIT DU VICOMTE DE MELVILLE (DVM)	41
OUEST DE LA BAIE D'HUDSON (OBH).....	44

Termes utilisés dans le tableau de situation du CTOB pour 2024

1. But

Aux termes de son mandat, le Comité technique de l'ours blanc (CTOB) présentera un rapport annuel au Comité administratif sur l'ours blanc (CAOB) sur la situation de chacune des 13 sous-populations canadiennes d'ours blancs, en s'appuyant sur les meilleures données scientifiques disponibles et sur le savoir autochtone (SA).

Le présent document définit les divers termes utilisés dans le tableau de situation et précise la base sur laquelle la situation de chaque sous-population a été évaluée par le CTOB en février 2024.

2. Définitions

2.1 Sous-population

La colonne « Sous-population » présente les noms des sous-populations d'ours blancs.

2.2 Estimation la plus récente de la population (année de l'estimation)

La colonne « Estimation la plus récente de la population (année de l'estimation) » présente les estimations scientifiques les plus récentes de l'abondance telles qu'établies par le CTOB. La période visée par chaque estimation est indiquée entre parenthèses.

2.3 ±2 ET ou IC à 95 %

La colonne « ±2 ET ou IC à 95 % » présente les intervalles de confiance (ou intervalles de crédibilité) des estimations les plus récentes de la population. Voir la référence concernant l'estimation de la population pour obtenir plus de détails sur le calcul d'un intervalle donné.

2.4 Méthode

La colonne « Méthode » présente les méthodes de relevé utilisées pour obtenir les estimations scientifiques les plus récentes de la population. Les abréviations suivantes sont utilisées pour désigner les méthodes de relevé : MR/P – relevé par marquage-recapture physique; MR/G – relevé par marquage-recapture génétique; A – relevé aérien.

2.5 Tendance historique (fondée sur la science)

La colonne « Tendance historique (fondée sur la science) » présente l'évaluation scientifique par le CTOB des changements dans l'abondance que les sous-populations ont connus depuis la signature en 1973 de l'Accord sur la conservation des ours blancs (polaires), laquelle a mené aux pratiques de gestion et aux recherches actuelles. L'estimation la plus récente de la population et la première estimation historique étayée qui soit comparable sont examinées pour chaque sous-population. Si une comparaison directe des estimations de l'abondance n'est pas possible ou s'il n'existe qu'une seule estimation de l'abondance, d'autres sources de données peuvent être utilisées pour effectuer cette évaluation (voir le point 3.1 « Étapes de la détermination des tendances historiques [fondées sur la science] »).

2.6 Tendance la plus récente (fondée sur la science)

La colonne « Tendance la plus récente (fondée sur la science) » présente l'évaluation par le CTOB de la direction la plus récente du changement de l'abondance qu'a connu une sous-population donnée. Cette évaluation vise à informer le CAOBS quant à l'augmentation, au déclin ou à la stabilité d'une sous-population donnée dans un passé récent, selon les meilleures données scientifiques disponibles. La tendance la plus récente (fondée sur la science) est évaluée au moyen d'une comparaison entre l'estimation la plus récente de la population (pourvu qu'elle se situe dans les 15 dernières années) et l'estimation précédente. Si la comparaison directe des estimations de la population est impossible ou non applicable, d'autres données, par exemple les données issues d'analyses de viabilité des populations (AVP), les indicateurs de productivité et la pression de récolte récente, peuvent être utilisées pour inférer tout changement récent de l'abondance. En outre, s'il y a eu plus d'une évaluation dans les 15 dernières années, toutes les évaluations devraient être prises en compte (voir le point 3.3 « Étapes de la détermination des tendances les plus récentes [fondées sur la science] »).

2.7 Tendance (fondée sur le savoir autochtone)

La colonne « Tendance (fondée sur le SA) » présente l'évaluation de la tendance d'une sous-population donnée d'ours blancs selon le SA, s'il est accessible. Cette évaluation est éclairée par l'expérience vécue et la mémoire vivante des détenteurs du SA (voir le point 4.1 « Étapes de la détermination des tendances [fondées sur le savoir autochtone] »).

Le Comité applique une définition du SA semblable à celle des connaissances écologiques traditionnelles (CET) adoptée par les États de l'aire de répartition de l'ours blanc. Le SA se rapporte à d'autres concepts similaires, notamment le savoir traditionnel et l'Inuit Qaujimagatuqangit.

Le savoir autochtone (SA) désigne un ensemble cumulatif de connaissances sur les relations des êtres vivants entre eux et avec leur environnement, qui est issu des pratiques culturelles, des expériences vécues et des traditions des peuples autochtones.

2.8 Prélèvement annuel (moyenne sur 5 ans)

La colonne « Prélèvement annuel (moyenne sur 5 ans) » présente la moyenne des prélèvements annuels déclarés au cours des cinq années précédentes, y compris tous les cas de mortalité anthropique connus et les prélèvements destinés aux zoos. Les peuples autochtones de l'Ontario et du Québec récoltent des ours blancs, mais la déclaration demeure volontaire, et aucun système de quotas n'est actuellement en place. Le Manitoba n'autorise pas la récolte, mais des prélèvements anthropiques (p. ex. prélèvements d'ours à problèmes ou destinés aux zoos) sont effectués. Les prélèvements déclarés par ces trois administrations sont inclus dans la colonne « Prélèvement annuel (moyenne sur 5 ans) ».

2.9 Prélèvement annuel (2021-2022)

La colonne « Prélèvement annuel (2021-2022) » présente les prélèvements déclarés au cours de la saison de récolte précédente, y compris tous les cas de mortalité anthropique connus et les prélèvements destinés aux zoos. Les peuples autochtones de l'Ontario et du Québec récoltent des ours blancs, mais la déclaration demeure volontaire, et aucun système de quotas n'est actuellement en place. Le Manitoba n'autorise pas la récolte, mais des prélèvements anthropiques (p. ex. prélèvements

d'ours à problèmes ou destinés aux zoos) sont effectués. Les prélèvements déclarés par ces trois administrations sont inclus dans la colonne « Prélèvement annuel (2021-2022) ».

2.10 Prélèvement maximal possible (2021-2022)

La colonne « Prélèvement maximal possible (2021-2022) » présente le nombre total annuel de cas de mortalité anthropique d'ours blancs dans une sous-population donnée qui sont autorisés selon les quotas, la récolte totale autorisée, les prises totales autorisées ou des ententes volontaires pour la saison de chasse précédente. En l'absence d'une limite de récolte établie dans une administration donnée, l'abréviation de cette administration est indiquée plutôt qu'une valeur précise. Le prélèvement maximal possible total est réparti par administration ou par région géographique. En ce qui concerne le Nunavut, où le Harvest Administration and Credit Calculation System (HACCS) est utilisé pour gérer la récolte d'ours blancs, la récolte totale autorisée est répartie entre l'allocation de base et les crédits demandés pour la dernière saison de chasse.

2.11 Commentaires

La colonne « Commentaires » présente des commentaires pertinents au sujet de la sous-population d'ours blancs en question. De plus amples renseignements sur chaque sous-population sont fournis dans la section « Descriptions des sous-populations ».

2.12 Administration(s)

La colonne « Administration(s) » présente la ou les administrations territoriales, provinciales et étrangères qui chevauchent l'étendue géographique de la sous-population en question.

3. Évaluation des tendances historiques et des tendances les plus récentes (fondées sur la science)

3.1 Étapes de la détermination des tendances historiques (fondées sur la science)

Il s'agit de comparer l'estimation la plus récente de la population à sa première estimation historique étayée qui soit comparable. Lorsque l'estimation la plus récente est directement comparable à une estimation historique, une désignation sans qualificatif (c.-à-d. déclin, stabilité ou augmentation) peut être utilisée.

Si l'estimation la plus récente n'est pas directement comparable à une estimation historique en raison de différences dans la zone d'étude ou la méthode employée, une comparaison peut tout de même être faite, mais toute évaluation des changements de l'abondance doit être inférée. En tel cas, un qualificatif est requis (c.-à-d. déclin probable, stabilité probable ou augmentation probable).

Lorsque les estimations de la population ne sont pas comparables ou que leur comparaison ne permet pas d'établir une différence avec un degré acceptable de certitude statistique, d'autres données, comme les plus récentes caractéristiques démographiques de la sous-population (p. ex. la structure par âge), peuvent être utilisées pour inférer les changements de l'abondance qu'a connus la sous-population. Les sources de données ne comprennent pas le SA. Ici également, un qualificatif est requis (c.-à-d. déclin probable, stabilité probable ou augmentation probable). Les autres sources de données doivent être justifiées dans les commentaires ainsi que la section « Descriptions des sous-populations ».

Deux étapes servent à évaluer le degré de certitude statistique. Premièrement, nous nous appuyons directement sur des rapports ou articles publiés. Si les auteurs indiquent expressément une différence

statistiquement certaine ou une absence de différence entre les deux estimations de la population, alors la tendance est présentée sans modification. En l'absence d'une telle évaluation, le CTOB a recours à son expertise technique collective pour évaluer le degré de certitude statistique. Dans les deux cas, une justification est fournie dans les commentaires ainsi que dans la section « Descriptions des sous-populations ».

Lorsque les données sont insuffisantes ou que les données disponibles ne sont pas assez fiables pour évaluer le changement de l'abondance qu'a connu une sous-population, sa tendance est évaluée comme étant incertaine.

Des renseignements supplémentaires sont fournis dans la colonne « Commentaires » du tableau de situation, notamment des renseignements relatifs à l'inscription, comme les principales menaces, et les autres sources de données qui pourraient avoir été utilisées.

3.2 Désignation des tendances historiques (fondées sur la science)

Déclin	L'estimation la plus récente de la population est plus basse que son estimation historique avec un degré acceptable de certitude statistique.
Stabilité	L'estimation la plus récente de la population n'est pas différente de son estimation historique.
Augmentation	L'estimation la plus récente de la population est plus haute que son estimation historique avec un degré acceptable de certitude statistique.
Déclin probable	L'effectif le plus récent de la population est inféré comme étant inférieur à son effectif historique.
Stabilité probable	L'effectif le plus récent de la population est inféré comme n'étant pas différent de son effectif historique.
Augmentation probable	L'effectif le plus récent de la population est inféré comme étant supérieur à son effectif historique.
Incertaine	Les données sont insuffisantes ou les données disponibles ne sont pas assez fiables pour évaluer la tendance de la population.

3.3 Étapes de la détermination des tendances les plus récentes (fondées sur la science)

Il s'agit de comparer l'estimation la plus récente de la population à son estimation précédente, pourvu que l'estimation la plus récente se situe dans les 15 dernières années. Lorsque l'estimation la plus récente de la population est directement comparable à son estimation précédente, une désignation sans qualificatif (c.-à-d. déclin, stabilité ou augmentation) peut être utilisée. En outre, s'il y a eu plus d'une évaluation dans les 15 dernières années, toutes les évaluations peuvent être prises en compte.

Si l'estimation la plus récente de la population n'est pas directement comparable à son estimation précédente en raison de différences dans la zone d'étude ou la méthode employée ou de la désuétude de l'estimation précédente, une comparaison peut tout de même être faite; cependant, toute évaluation des changements de l'abondance doit être inférée et un qualificatif est requis (c.-à-d. déclin probable, stabilité probable ou augmentation probable).

En l'absence d'un degré acceptable de certitude statistique quant à la différence entre les deux estimations de la population, ou lorsque ces estimations ne sont pas comparables ou ne permettent pas d'évaluer la tendance récente, d'autres données susceptibles d'offrir un aperçu de la situation de la population (p. ex. la répartition par âge ou l'état corporel) peuvent être utilisées pour inférer tout changement de l'effectif de la sous-population. Les sources de données ne comprennent pas le SA. Ici également, un qualificatif est requis (c.-à-d. déclin probable, stabilité probable ou augmentation probable).

Deux étapes servent à évaluer le degré de certitude statistique. Premièrement, nous nous appuyons directement sur des rapports ou articles publiés. Si les auteurs indiquent expressément une différence statistiquement certaine ou une absence de différence entre les deux estimations de la population, alors la tendance est présentée sans modification. En l'absence d'une telle évaluation, le CTOB a recours à son expertise technique collective pour évaluer le degré de certitude statistique. Dans les deux cas, une justification est fournie dans les commentaires ainsi que dans la section « Descriptions des sous-populations ».

Lorsque les données sont insuffisantes ou que les données disponibles ne sont pas assez fiables pour évaluer les changements de l'effectif de la population, ou s'il n'y a eu aucune estimation de la population dans les 15 dernières années, la tendance de la sous-population est évaluée comme étant incertaine.

Des renseignements supplémentaires sont fournis dans la colonne « Commentaires » du tableau de situation, notamment des renseignements relatifs à l'inscription, comme les principales menaces, et les autres sources de données qui pourraient avoir été utilisées.

3.4 Désignation des tendances les plus récentes (fondées sur la science)

Déclin	L'estimation la plus récente de la population est plus basse que son estimation précédente avec un degré acceptable de certitude statistique.
Stabilité	L'estimation la plus récente de la population n'est pas différente de son estimation précédente.
Augmentation	L'estimation la plus récente de la population est plus haute que son estimation précédente avec un degré acceptable de certitude statistique.
Déclin probable	L'effectif le plus récent de la population est inféré comme étant inférieur à son effectif précédent.
Stabilité probable	L'effectif le plus récent de la population est inféré comme n'étant pas différent de son effectif précédent.
Augmentation probable	L'effectif le plus récent de la population est inféré comme étant supérieur à son effectif précédent.
Incertaine	Les données sont insuffisantes ou les données disponibles ne sont pas assez fiables pour évaluer la tendance de la population, ou il n'y a eu aucune estimation de l'effectif de la population dans les 15 dernières années.

4. Évaluation des tendances (fondées sur le savoir autochtone)

4.1 Étapes de la détermination des tendances (fondées sur le savoir autochtone)

Les Autochtones qui ont participé à des pratiques culturelles, expériences vécues, traditions et observations ayant permis de peaufiner et de parfaire leur expertise spécialisée sont les détenteurs du SA. Les détenteurs du SA qui possèdent une expertise sur l'ours blanc peuvent faire part de leurs observations directes sur les tendances des sous-populations d'ours blancs, lesquelles sont souvent communiquées oralement. Les connaissances des détenteurs du SA sont consignées dans des études exhaustives sur le SA, et celles axées sur l'ours blanc comprennent souvent des évaluations des caractéristiques démographiques de l'espèce par les détenteurs de ce savoir. L'expertise des détenteurs du SA peut également être répertoriée d'autres façons, par exemple dans des rapports de consultation sur la gestion de la faune, des rapports de surveillance communautaire, des procès-verbaux d'audiences publiques et des articles évalués par les pairs.

Les détenteurs du SA, les études exhaustives sur le SA et les autres rapports et procès-verbaux concernant le SA sont des sources d'information que le Comité consulte pour déterminer les tendances (fondées sur le savoir autochtone). Ces sources sont mentionnées dans les notes en bas du tableau de situation.

Lorsqu'il y a une cohérence entre les expériences et observations des détenteurs du savoir autochtone à l'égard d'une tendance de l'abondance d'ours blancs dans une région précise au cours d'une période donnée, cette tendance (p. ex. déclin, stabilité ou augmentation) est présentée comme tendance (fondée sur le SA) de la sous-population d'ours blancs où ces expériences et observations ont eu lieu.

S'il y a une incohérence entre les expériences vécues et observations des détenteurs du SA à l'égard d'une tendance de l'abondance des ours blancs, la tendance (fondée sur le SA) est « incertaine ». Si l'incohérence est de nature géographique (par exemple, les observations sur une sous-population faites dans une région sont différentes de celles faites dans une autre région), les différentes tendances peuvent être indiquées entre parenthèses.

S'il y a un manque de connaissances autochtones consignées sur une tendance de l'effectif d'une sous-population, la tendance (fondée sur le SA) est « inconnue ».

4.2 Désignation des tendances (fondées sur le savoir autochtone)

Déclin	Les détenteurs du SA ont systématiquement déterminé que la population d'ours blancs a diminué au cours d'une période donnée.
Stabilité	Les détenteurs du SA ont systématiquement déterminé que la population d'ours blancs n'a pas changé au cours d'une période donnée.
Augmentation	Les détenteurs du SA ont systématiquement déterminé que la population d'ours blancs a augmenté au cours d'une période donnée.
Incertaine	Les détenteurs du SA ont déterminé différentes tendances à l'égard d'une sous-population d'ours blancs au cours d'une période donnée.
Inconnue	Les connaissances autochtones diffusées publiquement par les détenteurs du SA sont insuffisantes pour évaluer la tendance.

Tableau de situation du CTOB pour 2024

Sous-population	Estimation la plus récente de la pop. (année de l'estimation)	±2 ET ou IC/ICr à 95 %	Méthode †	Tendance historique (fondée sur la science)	Tendance la plus récente (fondée sur la science)	Tendance (fondée sur le SA)	Prélèvement annuel (moyenne sur 5 ans)‡	Prélèvement annuel (2022-2023)‡	Prélèvement maximal possible (2022-2023)‡	Commentaires	Administration(s)#
Baie de Baffin (BB)	2 826 (2012-2013 ¹)	2 059-3 593	MR/G	Incertaine	Stabilité probable ³ (de 1997 à 2013)	Stabilité ²	140	114	160 (Nt : 80 + Gl. : 80)	La comparaison directe entre l'estimation précédente (1997) et l'estimation actuelle (2012-2013) est impossible pour les raisons suivantes : différences dans l'étendue géographique et la répartition des ours; diminution de la glace de mer; augmentation du temps passé sur la terre ferme; déclin de l'état corporel; réduction de la période de mise bas; augmentation du trafic maritime. Évaluation du risque lié à la récolte terminée. Au Nunavut, des crédits se sont accumulés depuis 2018-2019 et il y avait un total de 52 crédits (33 femelles et 19 mâles) après la saison de récolte de 2022-2023.	Gl., Nt
Détroit de Davis (DD)	2 015 (2017-2018 ⁴)	1 603-2 588	MR/G	Augmentation probable	Déclin probable ⁶ (de 2007 à 2018)	Augmentation ⁵	60,6	53	Qc + 76 (Nt : 61 + T.-N.-L. : 12 + Gl. : 3)	Possibilité d'une récolte élevée (sous-population actuellement gérée aux fins de réduction de la population au Nunavut; aucun quota n'est en place au Québec dans la RMN); le changement apporté en 2019-2020 au rapport des sexes utilisé aux fins de gestion de la récolte au Nunavut pourrait causer une réduction du taux de croissance de la population; diminution de la glace de mer. Au Nunavut, des crédits se sont accumulés depuis 2012-2013 et il y avait un total de 213 crédits (81 femelles et 132 mâles) après la saison de récolte de 2022-2023.	Gl., T.-N.-L., Nt, Qc
Bassin Foxe (BF)	2 585 (2009-2010 ⁷)	2 096-3 189	A	Stabilité	Stabilité ⁹ (de 1997 à 2010)	Augmentation ⁸	113,4	117	Qc + 145 (Nt : 123 + 22 cr.)	Diminution de la glace de mer; possibilité d'une récolte élevée (aucun quota n'est en place au Québec dans la RMN); le changement apporté en 2019-2020 au rapport des sexes utilisé aux fins de gestion de la récolte au Nunavut pourrait causer une réduction du taux de croissance de la population. Au Nunavut, des crédits se sont accumulés depuis plus de 20 ans et il y avait un total de 145 crédits (47 femelles et 98 mâles) après la saison de récolte de 2022-2023.	Nt, Qc
Golfe de Boothia (GB)	1 525 (2015-2017 ⁴⁰)	949-2 101	MR/G	Stabilité probable	Stabilité (de 2000 à 2017)	Augmentation ¹¹	66,7	66	82 (Nt : 82)	La modification actuelle et prévue de l'habitat pourrait avoir une incidence sur la productivité de l'écosystème; productivité saine; le changement apporté en 2019-2020 au rapport des sexes utilisé aux fins de gestion de la récolte au Nunavut pourrait causer une réduction du taux de croissance de la population; possibilité d'une augmentation de la navigation touristique estivale. Au Nunavut, des	Nt

Sous-population	Estimation la plus récente de la pop. (année de l'estimation)	±2 ET ou IC/ICr à 95 %	Méthode †	Tendance historique (fondée sur la science)	Tendance la plus récente (fondée sur la science)	Tendance (fondée sur le SA)	Prélèvement annuel (moyenne sur 5 ans)‡	Prélèvement annuel (2022-2023)‡	Prélèvement maximal possible (2022-2023)‡	Commentaires	Administration(s)#
										crédits se sont accumulés depuis 2022-2023 et il y avait un total de 16 crédits (10 femelles et 6 mâles) après la saison de récolte de 2022-2023.	
Bassin Kane (BK)	357 (2013-2014 ¹³)	221-493	MR/G	Déclin probable	Augmentation ¹⁵ (de 1997 à 2014)	Augmentation ¹⁴	7,0	7	14 (Nt : 5 + Gl. : 9)	Petite population; survie des mâles adultes : 0,87 et survie des femelles : 0,95; changements de l'état de la glace de mer (la glace de plusieurs années est devenue saisonnière); possibilité d'une réaction positive aux premières répercussions des changements climatiques et à la réduction de la récolte; évaluation du risque lié à la récolte terminée. Au Nunavut, des crédits se sont accumulés depuis 2018-2019 et il y avait un total de 25 crédits (10 femelles et 15 mâles) après la saison de récolte de 2022-2023.	Gl., Nt
Détroit de Lancaster (DL)	2 541 (1995-1997 ¹⁶)	1 759-3 323	MR/P	Stabilité probable	Incertaine ¹⁸	Augmentation ¹⁷	64,3	62	87 (Nt : 85 + 2 cr.)	Diminution de la glace de mer; possibilité d'une augmentation de la navigation touristique et commerciale estivale; possible création de l'aire marine nationale de conservation Tallurutiup Imanga; le changement apporté en 2019-2020 au rapport des sexes utilisé aux fins de gestion de la récolte au Nunavut pourrait causer une réduction du taux de croissance de la population; la réévaluation a débuté en 2021 et est toujours en cours. Au Nunavut, des crédits se sont accumulés depuis plus de 20 ans et il y avait un total de 213 crédits (123 femelles et 90 mâles) après la saison de récolte de 2022-2023.	Nt
Détroit de M'Clintock (DM)	716 (2014-2016 ⁴¹)	545-955	MR/G	Incertaine	Augmentation (de 2000 à 2016)	Stabilité ²⁰	9,6	8	21 (Nt : 21)	Changements de l'état de la glace de mer (la glace de plusieurs années est devenue annuelle); possibilité d'une augmentation de la navigation touristique estivale; le changement apporté en 2019-2020 au rapport des sexes utilisé aux fins de gestion de la récolte au Nunavut pourrait causer une réduction du taux de croissance de la population. Au Nunavut, des crédits se sont accumulés depuis 2021-2022 et il y avait un total de 21 crédits (10 femelles et 11 mâles) après la saison de récolte de 2022-2023.	Nt
Nord de la mer de Beaufort (NMB)	1 291 (2006 ²²)	Aucune estimation	MR/P	Stabilité probable	Incertaine	Stabilité ²³	23,4	17	77 (Nt : 6 + T.N.-O. : 71)	Changements de l'état de la glace de mer (la glace de plusieurs années est devenue annuelle); faible récolte en raison de l'état de la glace, qui ne permet pas les déplacements, et faible pression de récolte; nouvelle limite officiellement acceptée en 2013;	T.N.-O., Nt

Sous-population	Estimation la plus récente de la pop. (année de l'estimation)	±2 ET ou IC/ICr à 95 %	Méthode †	Tendance historique (fondée sur la science)	Tendance la plus récente (fondée sur la science)	Tendance (fondée sur le SA)	Prélèvement annuel (moyenne sur 5 ans)‡	Prélèvement annuel (2022-2023)‡	Prélèvement maximal possible (2022-2023)‡	Commentaires	Administration(s)#
										l'abondance est en cours de réévaluation. Dans les T.N.-O., la récolte potentielle autorisée est fondée sur une estimation de la population corrigée aux fins de gestion, soit 1 711 ours. Au Nunavut, des crédits se sont accumulés depuis plus de 20 ans et il y avait un total de 142 crédits (49 femelles et 93 mâles) après la saison de récolte de 2022-2023.	
Baie Norwegian (BN)	203 (1997 ²⁵)	115-291	MR/P	Incertaine	Incertaine ²⁷	Stabilité ²⁶	1,4	3	4 (Nt : 4)	Petite population isolée soumise à une faible pression de récolte; la réévaluation devait commencer en 2021, mais elle a été reportée. Au Nunavut, des crédits se sont accumulés depuis plus de 20 ans et il y avait un total de 44 crédits (29 femelles et 15 mâles) après la saison de récolte de 2022-2023.	Nt
Sud de la mer de Beaufort (SMB)	1 215 (2006 ²⁸)	Aucune estimation	MR/P	Incertaine	Incertaine	Stabilité ²⁹	21,8	35	56 (É.-U. : 35 + RDI : 21)	Détérioration de l'état corporel, baisse de la croissance et déclin démographique liés à la diminution de la glace de mer; la limite est de la sous-population a été ajustée en 2013-2014; le SA laisse entendre qu'à mesure que la glace de mer fondra dans le sud de la mer de Beaufort, l'aire de répartition de l'ours blanc se déplacera vers le nord; possibilité d'une augmentation de l'exploitation pétrolière/gazière en Alaska; en cours de réévaluation.	T.N.-O/Yn, É.-U.
Détroit du Vicomte de Melville (DVM)	235x (2014)	148-569 (ICr)	MR/P	Augmentation	Augmentation	Augmentation ³⁵	1,0	0	7 (Nt : 3 + T.N.-O : 4)	Changements à long terme de l'état de la glace de mer (la glace de plusieurs années est devenue annuelle); petite population éloignée; les travaux effectués sur le terrain de 2012 à 2014 montraient une augmentation de la proportion de mâles dans la région depuis 1989-1992. L'estimation la plus récente (2014) laisse entendre que cette sous-population s'est rétablie après une période antérieure de forte récolte et qu'elle se serait stabilisée. La survie des femelles adultes a peut-être diminué. Au Nunavut, des crédits se sont accumulés depuis 2005-2006 et il y avait un total de 10 crédits (5 femelles et 5 mâles) après la saison de récolte de 2022-2023.	T.N.-O, Nt
Sud de la baie d'Hudson (SBH)	1 119 (2021)	860-1 454	A	Stabilité probable	Stabilité probable ³¹ (de 2011-2012 à 2021)	Stabilité dans la baie James; augmentation probable dans l'est de la baie d'Hudson ³²	40	33	Ont. + Qc + 48 (Nt : 25 + RMN : 23)	Augmentation de l'abondance entre les relevés de 2016 et de 2021 en raison de déplacements temporaires depuis l'ouest de la baie d'Hudson et/ou de l'amélioration des indices vitaux; réduction de l'habitat de mise bas sur le pergélisol entre 1980 et 2010; déclaration incomplète des cas de mortalité anthropique; le changement apporté en 2019-2020	Nt, Ont., Qc

Sous-population	Estimation la plus récente de la pop. (année de l'estimation)	±2 ET ou IC/ICr à 95 %	Méthode †	Tendance historique (fondée sur la science)	Tendance la plus récente (fondée sur la science)	Tendance (fondée sur le SA)	Prélèvement annuel (moyenne sur 5 ans)‡	Prélèvement annuel (2022-2023)‡	Prélèvement maximal possible (2022-2023)‡	Commentaires	Administration(s)‡
Ouest de la baie d'Hudson (OBH)	618 (2021)	425-899	A	Déclin probable	Déclin probable ³⁹ (de 2011 à 2021)	Augmentation ³⁸	33,8	37	Man. + 38 (Nt : 38)	<p>au rapport des sexes utilisé aux fins de gestion de la récolte au Nunavut pourrait causer une réduction du taux de croissance de la population. Au Nunavut, des crédits se sont accumulés depuis plus de 20 ans et il y avait un total de 5 crédits (5 femelles) après la saison de récolte de 2022-2023.</p> <p>Diminution de l'abondance entre les relevés de 2016 et de 2021 en raison de déplacements temporaires vers le sud de la baie d'Hudson et/ou d'indices vitaux plus faibles; productivité plus faible que celle des sous-populations adjacentes du bassin Foxe et du sud de la baie d'Hudson; lien entre la survie des femelles et l'état de la glace de mer; le changement apporté en 2019-2020 au rapport des sexes utilisé aux fins de gestion de la récolte au Nunavut pourrait causer une réduction du taux de croissance de la population. Au Nunavut, des crédits se sont accumulés depuis 2017-2018 et il y avait un total de 25 crédits (2 femelles et 23 mâles) après la saison de récolte de 2022-2023.</p>	Man., Nt

Notes

† Abréviations désignant les méthodes de relevé : A – relevé aérien; MR/P – relevé par marquage-recapture physique; MR/G – relevé par marquage-recapture génétique.

‡ Les valeurs des prélèvements effectués dans les cinq dernières années et dans la dernière année comprennent tous les cas déclarés de mortalité anthropique et tous les prélèvements effectués dans les administrations, qu'elles disposent ou non de quotas annuels établis par des conseils de gestion des ressources fauniques. Les récoltes non déclarées effectuées dans les administrations où la déclaration n'est pas obligatoire ne sont pas prises en compte dans ces valeurs.

Abréviations désignant les entités administratives : Gl. – Groenland; Man. – Manitoba; T.-N.-L. – Terre-Neuve-et-Labrador; Nt. – Nunavut (récolte totale autorisée de base + crédits demandés - toute réduction de l'allocation de base); T.N.-O. – Territoires du Nord-Ouest; Ont. – Ontario; Qc – Québec; É.-U. – États-Unis; Yn – territoire du Yukon; RDI – région désignée des Inuvialuit; RMN – région marine du Nunavik.

* Le « sud de la baie d'Hudson » ne respecte pas l'ordre alphabétique. La ligne en question a été déplacée au-dessus de celle présentant la sous-population adjacente de l'ouest de la baie d'Hudson de façon à mettre en évidence les déplacements possibles entre les deux sous-populations.

Notes supplémentaires

1. SWG (2016).
2. Born *et al.* (2011); Dowsley (2005); Dowsley (2007); Dowsley et Taylor (2006a); procès-verbaux et mémoires des audiences publiques du Conseil de gestion des ressources fauniques du Nunavut (CGRFN), avril 2008 et septembre 2009.
3. SWG (2016).
4. Dyck *et al.* (2022) : rapport final préparé pour le ministère de l'Environnement du gouvernement du Nunavut.
5. Kotierk (2010a, 2010b); York *et al.* (2015); Tomaselli *et al.* (2022).
6. Peacock *et al.* (2013); Stirling *et al.* (1980); Dyck *et al.* (2022).
7. Stapleton *et al.* (2016).
8. Canadian Wildlife Service (2009) : *Nunavut Consultation Report*; Dyck, comm. pers. 7 février 2013; Sahanatien, comm. pers. 7 février 2013; Tomaselli *et al.* (2022).
9. Stapleton *et al.* (2016); Taylor *et al.* (2006b).
10. Taylor *et al.* (2009).
11. Canadian Wildlife Service (2009) : *Nunavut Consultation Report*; Keith *et al.* (2005); Wong *et al.* (2021) : *Inuit Qaujimagatunangit of Gulf of Boothia polar bears – Final report*.
12. Les indices vitaux datent de 2000 (Taylor *et al.*, 2009) et sont considérés comme trop vieux/peu fiables pour l'AVP.
13. SWG (2016).
14. Canadian Wildlife Service (2009) : *Nunavut Consultation Report*.
15. SWG (2016).

16. Schweinsburg *et al.* (1980); Taylor *et al.* (2008).
17. Canadian Wildlife Service (2009) : *Nunavut Consultation Report*.
18. Selon le régime de récolte historique (de 1993 à 1997), la population devait être stable de 1997 à 2012. D'après le taux de récolte moyen de 78 ours par année (de 2002 à 2006), et en nous fondant sur une AVP, nous estimons que la population est plus susceptible de diminuer que d'augmenter (Taylor *et al.*, 2008). Le taux de récolte actuel devrait également entraîner un déclin, mais aucun indice vital récent n'a été calculé en vue de mettre à jour l'AVP.
19. Taylor *et al.* (2006a).
20. Les Inuits signalent que partout dans la région, les ours se déplacent vers des régions avoisinantes (Canadian Wildlife Service [2009] : *Nunavut Consultation Report*; Keith *et al.*, 2005); Wong *et al.* (2021) : *Inuit Qaujimagatuqangit of M'Clintock Channel polar bears – Final report*.
21. Il s'agit probablement d'une augmentation, d'après une évaluation quantitative du taux de croissance (Taylor *et al.*, 2006a).
22. Griswold *et al.* (2017); Stirling *et al.* (2011).
23. Joint Secretariat (2015).
24. La taille de population utilisée aux fins de gestion a été ajustée à 1 200 en raison d'un biais dans l'estimation de la population (Amstrup *et al.*, 2005; Stirling *et al.*, 2011).
25. Taylor *et al.* (2006a); Taylor *et al.* (2008).
26. Canadian Wildlife Service (2009) : *Nunavut Consultation Report*.
27. Les indices vitaux utilisés pour l'AVP de Riskman datent d'il y a 20 ans et ont été remplacés par ceux d'autres populations (Taylor *et al.*, 2008); aucune étude récente n'a été effectuée dans la région.
28. Bromaghin *et al.* (2015); Griswold *et al.* (2017); USFWS (2010).
29. Joint Secretariat (2015).
30. L'estimation de la population est plus basse que ses estimations précédentes, mais elle n'est pas statistiquement différente de celles-ci (Amstrup *et al.*, 1986; Regehr *et al.*, 2006). Les quotas étaient fondés sur la supposition que la récolte totale de femelles indépendantes ne dépassera pas le maximum durable modélisé, soit 1,5 % de la population (Taylor *et al.*, 1987), et qu'un rapport de 2 mâles pour 1 femelle sera maintenu dans la récolte totale (Stirling, 2002).
31. Northrup *et al.* (2022).
32. Étude sur le savoir inuit du Conseil de gestion des ressources fauniques de la région marine du Nunavik (CGRFRMN), 2018; procès-verbal de l'audience publique du CGRFRMN à Inukjuak, février 2014.
33. D'après une comparaison avec les estimations précédentes de la sous-population (Obbard *et al.*, 2018; Obbard *et al.*, 2016; Obbard *et al.*, 2013; Obbard, 2008; Kolenosky, 1994).
34. Taylor *et al.* (2002).
35. *Nunavut Consultation Report* de 2009, préparé par le Service canadien de la faune; consultations communautaires de 2012 et de 2013.
36. Depuis le dernier relevé, la récolte a fait l'objet d'une gestion aux fins de croissance de la population, y compris un moratoire de cinq ans; les portées étaient de taille comparable en 2012 (GNWT, données inédites).
37. Dyck *et al.* (2017); voir l'estimation par marquage-recapture de Lunn *et al.* (2016).
38. *Nunavut Consultation Report* de 2009, préparé par le Service canadien de la faune; Kotierk (2012); Nirlungayuk et Lee (2009); procès-verbaux des audiences publiques du CGRFN, 2005, 2011, 2014 et 2017; Tyrrell (2006).
39. Stapleton *et al.* (2014); Lunn *et al.* (2016); Dyck *et al.* (2017).
40. Dyck *et al.* (2020a).
41. Dyck *et al.* (2020b).

2024 PBTC Removals Table

2017/2018 – 2021/2022 RECORDED HUMAN-CAUSED MORTALITY IN POLAR BEAR POPULATIONS IN AND SHARED WITH CANADA COMPILED FOR 2023 PBTC MEETING

Population	Jurisdiction	2018-2019				2019-2020				2020-2021				2021-2022				2022-2023				5 year mean	3 year mean	Most recent year		N est.*	
		KILL	F	M	U	KILL	F	M	U	KILL	F	M	U	KILL	F	M	U	KILL	F	M	U	Harvest	Harvest	Harvest	Proportion Female		
BB	NU	73	34	39	0	74	32	42	0	68	22	46	0	65	32		33	0	66	34	32	0	140.0	135.0	114	0.4211	2826
	GL	78	32	46	0	70	26	44	0	77	24	53	0	81	28		51	2	48	14	34	0					
DS	NU	31	10	21	0	50	20	30	0	38	17	21	0	33	18		15	0	33	14	19	0	60.6	53.0	48	0.4043	2015
	NL	10	4	6	0	12	4	8	0	7	0	7	0	12	3		9	0	13	4	8	1					
	QC	21	2	19	0	17	10	7	0	5	1	3	1	10	2		8	0	2	1	1	0					
	GL	0	0	0	0	3	1	0	2	5	1	2	2	1	0		1	0	0	0	0	0	0				
FB	NU	114	42	72	0	109	39	70	0	111	54	57	0	109	47		62	0	117	49	68	0	113.4	112.3	117	0.4188	2585
	QC	7	4	3	0	0	0	0	0	0	0	0	0	0	0		0	0	0	0	0	0					
GB	NU	66	22	44	0	70	25	45	0	68	29	39	0	66	21		45	0	66	30	36	0	67.2	66.7	66	0.4545	1525
KB	NU	0	0	0	0	0	0	0	0	0	0	0	0	0	0		0	0	0	0	0	0	7.0	7.0	0	0.5714	357
	GL	4	2	2	0	10	3	7	0	6	1	5	0	8	3		5	0	7	4	3	0					
LS	NU	80	25	55	0	70	24	46	0	56	21	35	0	73	18		55	0	62	32	30	0	68.6	64.3	62	0.5161	2541
MC	NU	11	5	6	0	7	1	6	0	9	3	6	0	13	5		8	0	8	6	2	0	9.6	10.0	8	0.7500	716
NB	NWT	33	12	20	1	25	7	17	1	16	10	6	0	24	13		10	1	17	8	9	0	23.4	19.0	17	0.4706	1291
	NU	2	1	1	0	0	0	0	0	0	0	0	0	0	0		0	0	0	0	0	0					
NW	NU	0	0	0	0	0	0	0	0	0	0	0	4	0		4	0	3	2	1	0	1.4	2.3	3	0.6667	203	
SB	NWT/YK	8	3	5	0	0	0	0	0	0	0	0	0	0	0		0	0	5	2	3	0	21.8	23.0	35	0.3077	1215
	USA	18	8	5	5	15	2	12	1	13	4	9	0	21	4		12	5	30	6	15	9					
SH	NU	23	8	15	0	23	6	17	0	47	19	28	0	31	8		23	0	25	9	16	0	40	42.0	33	0.3667	1119
	ON	5	1	0	4	5	2	0	3	1	1	0	0	0	0		0	0	4	0	1	3					
	QC	11	2	9	0	7	3	4	0	7	2	3	2	6	1		2	3	4	2	2	0					
VM	NWT	0	0	0	0	0	0	0	0	0	0	0	0	0	0		0	0	0	0	0	0	0.8	1.0	0	0	161
	NU	1	0	1	0	0	0	0	0	3	1	2	0	0	0		0	0	0	0	0	0					
WH	NU	37	19	18	0	19	3	16	0	40	15	25	0	33	14		19	0	34	18	16	0	33.8	36.7	37	0.5405	618
	MB**	2	1	1	0	1	1	0	0	0	0	0	0	0	0		0	0	3	2	1	0					

Notes

Statistics represent a harvest season (July 1 to June 30)

% female calculation excludes bears of unknown sex

* Population estimate numbers taken from 2023 PBTC Status Table

** Includes live removals

Red text indicates changes to past years' numbers after DNA verification of reported kills or updated reported numbers; 21/22 numbers are subject to change

2024 PBTC Polar Bear Subpopulation Narratives

Baffin Bay (BB)

Status and Delineation

Based on movements of adult females with satellite radio-collars and recaptures of tagged animals, the Baffin Bay subpopulation is bounded by the North Water Polynya to the north, Greenland to the east and Baffin Island, Canada to the west (Taylor and Lee 1995; Taylor *et al.* 2001; Laidre *et al.* 2013, 2018). A distinct southern boundary at Cape Dyer on Baffin Island in Nunavut, Canada is evident from the movements of tagged bears (Stirling *et al.* 1980; Peacock *et al.* 2012) and from polar bears monitored by satellite telemetry (Taylor *et al.* 2001; Laidre *et al.* 2018). This boundary overlaps with the northern boundary of the Davis Strait subpopulation. Studies of microsatellite genetic variation have not revealed significant differences between polar bears in BB and neighboring Kane Basin, although there was significant genetic variation between polar bears in BB and those in Davis Strait (Paetkau *et al.* 1999; Peacock *et al.* 2015; Malenfant *et al.* 2016; SWG 2016). However, polar bears in BB cluster with bears in northern Davis Strait (Peacock *et al.* 2015).

An initial subpopulation estimate of 300 – 600 bears in BB was based on mark-recapture data collected in spring (1984 – 1989) in which the capture effort was restricted to shore-fast ice and the floe edge off northeast Baffin Island. However, work in the early 1990s showed that an unknown proportion of the subpopulation was typically offshore during the spring and, therefore, unavailable for capture. A second study (1993 – 1997) was carried out during September and October, when all polar bears were thought to be ashore in summer retreat areas on Bylot and Baffin islands (Taylor *et al.* 2005). Taylor *et al.* (2005) estimated the number of polar bears in BB at $2,074 \pm 226$ (SE).

Before the introduction of a quota system in Greenland during 2006, it was believed that the combined Nunavut/Greenland harvest removals for BB were not sustainable. A 2004 computer PVA simulation estimated that the BB subpopulation abundance was around 1600 bears (Aars *et al.* 2006), and that the population was likely declining. As a response to this, the Canada-Greenland Joint Commission (JC) on Polar Bear was established in 2009 with the objectives to manage polar bears within Kane Basin and Baffin Bay and to ensure their conservation, as well as establishing effective management systems (SWG 2016). Due to the high uncertainty of the population status, which was caused in part by harvest uncertainties and changing environmental conditions, a phased-in reduction in harvest levels was recommended by the JC between 2010 and 2014 where Nunavut harvest levels were reduced from 105 bears annually in 2010, to 65 bears annually in 2014. One primary objective of the JC was also to conduct a new population study to up-date the status of the BB subpopulation.

A 3-year genetic mark-recapture survey (via biopsy darting) was completed in 2014 resulting in a new population estimate, survival rates, and habitat use analyses (SWG 2016, Atkinson *et al.* 2021). The mean estimate of total abundance of the BB subpopulation in 2012-2013 was 2,826 (95% CI = 2,059-3,593) polar bears. Due to evidence that the sampling design and environmental conditions resulted in an underestimate of abundance in the 1990s, these two estimates are not directly comparable and trend in abundance cannot be determined.

Satellite telemetry data and habitat selection studies in the 2000s indicate a number of ecological changes related to sea ice loss in Baffin Bay. There has been a significant reduction in the range of the subpopulation in all months and seasons when compared to the 1990s. The most marked reduction is a 60% decline in subpopulation range size in summer. Emigration from Baffin Bay has declined since the 1990s, especially with a reduction of bears moving from BB into Davis Strait and Lancaster Sound. The total number of bears marked during studies in 2011-2012 in BB was equivalent to ~34% of the estimated population size. Despite this, instances of emigration were $\leq 1\%$ of the recaptures and recoveries of marks for the BB subpopulation.

Compared to the 1990s, adult female BB bears now use significantly lower sea-ice concentrations in winter and spring and spend 20-30 days longer on land on Baffin Island during the summer ice-free season (Laidre *et al.* 2018). Changes in maternity denning have been observed; entry dates into maternity dens are >1 month later in the 2000s than the 1990s. Furthermore, the first date of arrival on land of pregnant females is significantly earlier in the 2000s. Maternity dens in the 2000s occurred at higher elevations and steeper slopes than the 1990s, likely due to reduced snow cover (Escajeda *et al.* 2018).

Harvest Management

BB polar bears are harvested by hunters from Nunavut and Greenland. A harvest risk assessment of the BB polar bear subpopulation was completed for the JC by the SWG to provide various harvest scenarios to guide management decisions (Regehr *et al.* 2017). The harvest risk analyses incorporated various demographic approaches and vital rates based on polar bear life history, and potential effects of future sea-ice conditions on polar bear population size and status through projected trends in carrying capacity. The JC decided on a low-to-medium risk tolerance for the BB polar bear subpopulation with a management objective of maintaining a subpopulation size that is in balance with the number of bears the environment can support. The new harvest levels for BB represent a total removal rate of 5.7%, or 160 bears per year, at an overall sex ratio of 1 male to 1 female, and which is divided evenly between Canada and Greenland.

Protected Areas

Auyuittuq and Sirmilik National Parks on eastern Baffin Island, Canada attract visitors mainly for hiking and climbing. Both parks provide denning and summer habitat protection for BB polar bears. In Greenland, the Melville Bay Nature Reserve was created in 1980 in part to protect denning habitat. The reserve consists of an outer zone, where hunting is permitted and an inner zone in which all hunting is prohibited.

Indigenous Knowledge

Dowsley and Wenzel (2008) collected IK of the BB subpopulation through semi-directed interviews. While they found significant differences in the responses among communities in Nunavut regarding whether there had been any change in the size of the subpopulation and numbers of bears in town, the majority of respondents in each community reported an increase. No respondent indicated a decrease in the subpopulation or a decrease in the numbers of bears in communities. Respondents also observed receding of the floe-edge towards land, and a decrease in the amount of land-fast ice in the region. Comments from the public at the Nunavut Wildlife Management Board (NWMB) April 2008 public meeting, including hunters and trappers and Nunavut Tunngavik Incorporated (NTI) support the findings of Dowsley and Wenzel (2008).

In Greenland, an interview study carried out in 2006 (Born *et al.* 2011) reported an increased occurrence of polar bears close to the coast, polar bears being thinner, a decrease in sea ice cover and warmer, more unpredictable weather that influenced travelling and hunting activities, as well as an increased number of polar bears caught from skiffs, as compared to dog-sled (hunting polar bears from skidoos is forbidden in Greenland).

References

- Aars, J., Lunn, N.J., and Derocher, A.E. (eds.). 2006. *Polar Bears: Proceedings of the 14th Working Meeting of the IUCN/SSC Polar Bear Specialist Group, 20-24 June 2005, Seattle, Washington*. Gland, Switzerland and Cambridge, UK.
- Atkinson, S.N., Laidre, K.L., Arnold, T.W., Stapleton, S., Regehr, E.V., Born, E.W., Wiig, Ø., Dyck, M., Lunn, N.J., Stern, H.L., and Paetkau, D. 2021. A novel mark-recapture-recovery survey using genetic sampling for polar bears *Ursus maritimus* in Baffin Bay. *Endangered Species Research* **46**:105–120.
- Born, E.W., Heilmann, A., Kielsen Holm, L., and Laidre, K. 2011. Polar Bears in Northwest Greenland: An Interview Survey About the Catch and the Climate (Monographs on Greenland – Man & Society) Volume 351, ISBN 9788763531689, 250 pp.
- Dowsley, M., and Wenzel, G.W. 2008. "The time of the most polar bears": A co-management conflict in Nunavut. *Arctic* **61**:177–189.
- Escajeda, E., Laidre, K.L., Born, E.W., Wiig, Ø., Atkinson, S., Dyck, M., Ferguson, S.H., and Lunn, N.J. 2018. Identifying shifts in maternity den phenology and habitat characteristics of polar bears (*Ursus maritimus*) in Baffin Bay and Kane Basin. *Polar Biology* **41**:87–100.
- Laidre, K.L., Born, E.W., Atkinson, S.N., Wiig, Ø., Anderson, L.W., Lunn, N.J., Dyck, M., Regehr, E.V., McGovern, R. and Heagerty, P. 2018. Range contraction and increasing isolation of a polar bear subpopulation in an era of sea ice loss. *Ecology and Evolution* **8**:2062–2075.
- Laidre, K.L., Born, E.W., Gurarie, E., Wiig, Ø., Dietz, R., and Stern, H. 2013. Females roam while males patrol: divergence in breeding season movements of pack ice polar bears (*Ursus maritimus*). *Proceedings of the Royal Society B* **280**:20122371, <https://doi.org/10.1098/rspb.2012.2371>.
- Malenfant, R.M., Davis, C.S., Cullingham, C.I., and Coltman, D.W. 2016 Circumpolar genetic structure and recent gene flow of polar bears: a reanalysis. *PLoS ONE* **11**(3):e0148967, <https://doi.org/10.1371/journal.pone.0148967>.
- Paetkau, D., Amstrup, S.C., Born, E.W., Calvert, W., Derocher, A.E., Garner, G.W., Messier, F., Stirling, I., Taylor, M.K., Wiig, Ø., and Strobeck, C. 1999. Genetic structure of the world's polar bear populations. *Molecular Ecology* **8**:1571–1584.
- Peacock, E., Laake, J., Laidre, K.L., Born, E.W., and Atkinson, S.N. 2012. The utility of harvest recoveries of marked individuals to assess polar bear (*Ursus maritimus*) survival. *Arctic* **65**:391–400.
- Peacock, E., Sonsthagen, S.A., Obbard, M.E., Boltunov, A., Regehr, E.V., Ovsyanikov, N., Aars, J., Atkinson, S.N., Sage, G.K., Hope, A.G., Zeyl, E., Bachmann, L., Ehrich, D., Scribner, K.T., Amstrup, S.C., Belikov, S., Born, E., Derocher, A.E., Stirling, I., Taylor, M.K., Wiig, Ø., Paetkau, D., and Talbot, S.L. 2015. Implications of the circumpolar genetic structure of polar bears for their conservation in a rapidly warming Arctic. *PLoS ONE* **10**:e112021.
- Regehr, E.V., Atkinson, S., Born, E.W., Laidre, K.L., Lunn, N.J., and Wiig, Ø. 2017. Harvest assessment for the Baffin Bay and Kane Basin polar bear subpopulations: Final Report to the Canada-Greenland Joint Commission on Polar Bear, 31. July 2017. 107 pp.
- Stirling, I., Calvert, W., and Andriashek, D. 1980. Population ecology studies of the polar bear in the area of southeastern Baffin Island. Canadian Wildlife Service Occasional Paper No. 44, 33 pp.
- SWG [Scientific Working Group to the Canada-Greenland Joint Commission on Polar Bear]. 2016. Re-Assessment of the Baffin Bay and Kane Basin Polar Bear Subpopulations: Final Report to the Canada-Greenland Joint Commission on Polar Bear. 31 July 2016: x + 636 pp.
- Taylor, M.K., and Lee, J. 1995. Distribution and abundance of Canadian polar bear populations - a management perspective. *Arctic* **48**:147–154.
- Taylor, M.K., Akeagok, S., Andriashek, D., Barbour, W., Born, E.W., Calvert, W., Cluff, H.D., Ferguson, S., Laake, J., Rosing-Asvid, A., Stirling, I., and Messier, F. 2001. Delineating Canadian and Greenland polar bear (*Ursus maritimus*) populations by cluster analysis of movements. *Canadian Journal of Zoology* **79**:690–709.
- Taylor, M.K., Laake, J., McLoughlin, P.D., Born, E.W., Cluff, H.D., Ferguson, S.H., Rosing-Asvid, A., Schweinsburg, R., and Messier, F. 2005. Demography and viability of a hunted population of polar bears. *Arctic* **58**:203–214.

Davis Strait (DS)

Status and delineation

Based on the recapture or harvest of previously tagged animals and tracking of adult female polar bears with satellite collars, the Davis Strait subpopulation is delineated in Canada within the Labrador Sea, eastern Hudson Strait, Davis Strait south of Cape Dyer, and along a portion of southwest Greenland (Stirling et al. 1980, Stirling and Kiliaan 1980, Taylor and Lee 1995, Taylor et al. 2001). Several studies have examined the genetic structure of DS bears compared to adjacent subpopulations and examined gene flow patterns to determine if genetic variability across individual polar bears corresponds to the currently recognized subpopulation boundaries (see Paetkau et al. 1999, Crompton et al. 2008, 2014, Peacock et al. 2013). While genetic data indicate gene flow occurs, Paetkau et al. (1999) concluded that the definition of a Management Unit (MU) was met in DS. Peacock et al. (2013) identified marked genetic differences between northern and southern DS bears and suggested that continued climate warming may increase that separation between those two groups of bears due to Hudson Strait ice dynamic. To date, the DS subpopulation boundary and management unit has been established based on western scientific studies and may not reflect Inuit Knowledge related to polar bear distribution and seasonal movements.

According to mark-recapture studies conducted between 1974 and 1979 on sea ice in spring, 700-900 bears were estimated to be present in the southern Baffin Island portion of the current delimitation of DS and 60-90 additional bears in the northern Labrador coast portion (Stirling et al. 1980, Stirling and Kiliaan 1980). In 1993 the PBTC established the DS subpopulation abundance estimate at 1,400 polar bears to account for the bias in sampling in the original studies. This estimate was subjectively raised again to 1,650 in 2005 based on the minimum population size that would be needed to sustain the harvest level occurring at that time and the fact that Indigenous knowledge suggested that more bears were being seen over the last 20 years. In addition, harp seals (*Pagophilus groenlandicus*), an important prey species for that population, had increased dramatically over the same period (DFO 2010), providing a much-enhanced potential prey base for polar bears.

Because of the uncertainties surrounding the population status, the Government of Nunavut (GN) conducted another population inventory from 2005 - 2007, resulting in an abundance estimate of 2,158 (95% CI: 1833 – 2542) bears (Peacock et al. 2013). At that time, the subpopulation was assessed as stable but was displaying lowered reproductive rates. Polar bear survival in DS varied with time and geography and was related to factors that included reductions in sea ice habitat and increases of harp seal numbers (Peacock et al. 2013). It was suggested that the observed lowered reproductive rates and the decline in body condition of polar bears in DS at the time of the 2005-2007 survey could be a result of habitat changes and/or increased polar bear density (Peacock et al. 2013, Rode et al. 2012).

A two-year genetic-mark-recapture (biopsy) study was conducted in 2017 and 2018 involving all of the DS management jurisdictions and Boards. The study design resembled the 2005-2007 study in terms of coverage and timing to allow for comparison with the earlier study. Furthermore, the similarity in study design and methods allowed for a reanalysis of the 2005-2007 study data in conjunction with the 2017-2018 study data to improve accuracy of estimates. The analysis included the previous 2005-2007 live-capture data, harvest recovery data from 2005 to 2018, and the recent genetic samples collected in 2017 and 2018. Using this data set, DS abundance was re-estimated for the 2005-2007 at 2,250 bears [95% CRI 1,989 - 2,512], which falls within the confidence interval of the estimated abundance published by Peacock et al. (2013). Estimated abundance for the 2017-2018 period was 2,015 bears [95% CRI 1,603 - 2,588]. Geometric mean subpopulation growth between 2006 and 2018 was 0.989 [95% CRI 0.974 – 1.010] which corresponds to a 0.896 probability that subpopulation growth was <1 and thus the subpopulation most likely declined over this period. Mean annual reported harvest from all jurisdictions combined increased from 64.1 ± 10.1 (SD) bears/year between 1999 and 2008 to 86.8 ± 23.6 between 2009 and 2019. The increased harvest may be a factor explaining the lower abundance estimate in 2017-2018. Bears were however less likely to be in poor body condition during the 2017-18 study period when compared to the 2005-07 study. Over the two study periods, mean cub-of-the-year (COY) recruitment (number of COYs per adult females) ranged from 0.23 to 0.45 and mean yearling recruitment (number of yearlings per adult female) ranged from 0.23 to 0.41, which appear to be sufficient to sustain the subpopulation. Survival rates estimated for all segments of the population were slightly lower than those calculated by Peacock et al. (2013), but fell within their confidence intervals. Inter-annual variations in survival did not appear to be linked to any environmental variables that were assessed, including sea ice parameters.

Indigenous Knowledge

In addition to the 2017-2018 scientific study, IK studies occurred in Nunavut, Nunavik, and Nunatsiavut. The Nunavik Marine Region Wildlife Board released the Davis Strait portion of their comprehensive study of Nunavik Inuit Knowledge and values in 2019 (NMRWB, 2019). The report includes current and historical information on polar bear health, feeding, abundance, distribution, and migration, as well as Inuit use of polar bears as an economic, cultural, and food resource. In 2015, the Torngat Wildlife & Plants Co-Management Board released a report on Labrador polar bear IK based on interviews with knowledge holders in Nunatsiavut. It covered polar bear health and physical condition, distribution and demography, abundance, their physical environment and changes to it, as well as hunting and conservation (York et al. 2015). In Nunavut, a collaborative research project with Inuit hunters occurred in Pangnirtung and Kimmirut in 2019. The study focused on the importance of polar bears to Inuit, uses of polar bear, health, abundance, demography, distribution, ecology, and habitat. All Inuit respondents from both Pangnirtung and Kimmirut reported that polar bears have increased in abundance since the 1940s to the 2010s. Many respondents also reported an increase in polar bear abundance since the 1980s, although two respondents from Kimmirut reported that they have observed a recent decline in abundance since the 2000s (Tomaselli et al. 2022).

Harvest management

The DS subpopulation is shared between Greenland, Nunavut, Nunatsiavut (Newfoundland and Labrador) and Nunavik (Québec). Nunavut has a quota of 61 bears, Greenland has a quota of 3, Nunatsiavut has a quota of 12; there is currently no quota in place in Nunavik.

Greenland

This region currently has a quota which was introduced in 2006 with mandatory reporting and monitoring.

Nunavut

This region has had a managed harvest with mandatory reporting for 40+ years. Community compliance and reporting is high (>98%) with Conservation Officers in each community to ensure proper monitoring of the polar bear harvest. The Total Allowable Harvest (TAH) was increased in 2012/2013 from 46 to 61 bears annually in order to slightly reduce the DS subpopulation. The average annual removal for the 2012/2013 – 2021/2022 time periods has however not increased with the higher TAH, averaging 42.8 bears/year. The Flexible Quota System (FQS), introduced in 1995/1996, involved a 2:1 male to female harvest sex ratio; however, after extensive consultations and public feedback from communities in Nunavut, and the development of the Nunavut Polar Bear Co-Management Plan, changes to the polar bear harvest administration were introduced in 2019/2020. With the new co-management plan, which was approved by the Nunavut Wildlife Management Board, the harvest sex ratio was changed where communities could harvest up to 1 female for every male. Due to the protections that are in place for denning bears and females with offspring (family groups) the potential that this management change could be the cause for a decline of the DS subpopulation is relatively low. Adult female polar bears are the most important contributors to population growth and the harvest sex ratio has remained close to a 2:1 male to female ratio since the switch to the “up to 1:1 harvest ratio”. The effects from the change in harvest sex ratio on the actual sex ratio of harvested bears will continue to be monitored by the Government of Nunavut. The average female proportion of harvest in 2019/2020 until 2021/2022 is 30.1%. Females have been underharvested relative to the annual recommended quota by approximately 38.2% on average during 2017/2018 – 2021/2022 (5 yr) time period. In February 2022, the Polar Bear Flexible Quota System was replaced by the Harvest Administration and Credit Calculation System. The up to 1:1 harvest sex-ratio is still in place.

Nunatsiavut

In Newfoundland and Labrador, polar bears are listed as vulnerable under the Provincial Endangered Species Act (ESA), and general direction on management objectives is provided under the auspices of the ESA. The Government of NL, in consultation with Nunatsiavut Government (NG), Parks Canada and the Torngat Wildlife & Plants Co-Management Board, are in the process of finalizing a polar bear management plan that will set out these general management objectives. Harvest management implementation will also be updated in accordance with the Labrador Inuit Land Claim Agreement (LILCA). Currently, the annual total allowable harvest (TAH) for Nunatsiavut is 12 bears of either sex, with no harvest of denning bears or females with young of the year permitted. For all bears harvested by Nunatsiavut beneficiaries, there is mandatory reporting with full compliance. Issues relating to inter-jurisdictional harvest inequity and reporting requirements remain key focal areas.

Québec

Currently, there is no quota or mandatory reporting in Québec. The exact number of bears harvested annually is unknown but reported harvest levels between 2012/2013 – 2021/2022 have ranged from 5 bears per year to 61, with an average annual removal of 25 bears. The incomplete registration of the harvest done within DS by Québec could represent a potential concern for this subpopulation since the actual total harvest in the subpopulation is currently unknown. The Québec government, in collaboration with the Nunavik and Eeyou Marine Regions Wildlife Boards, is in the process of adopting its polar bear management plan. This plan will provide a framework for subsequent harvest monitoring, reporting and the management of polar bear subpopulations occurring within Québec, including DS.

Protected areas

The Torngat Mountains National Park is in northern Labrador, which protects polar bear onshore and denning habitat from development.

References

- Crompton, A.E., Obbard, M.E., Petersen, S.D., and Wilson, P.J. 2008. Population genetic structure in polar bears (*Ursus maritimus*) from Hudson Bay, Canada: Implications of future climate change. *Biological Conservation* 141: 2528-2539.
- Department of Fisheries and Oceans Canada (DFO). 2010. Current status of Northwest Atlantic harp seal, *Phagophilus groenlandicus*. Science Advisory Report 2009/74. DFO Canadian Scientific Advisory Secretariat, Ottawa, Ontario, Canada.
- Nunavik Marine Regional Wildlife Board (NMRWB). 2019. Nunavik Inuit Knowledge and Observations of Polar Bears: Polar bears of the Davis Strait sub-population. Project conducted and report prepared for the NMRWB by Basterfield, M., Furgal, C., Breton-Honeyman, K., Rae, J, and O'Connor, M. 111 pp.
- Paetkau, D., Amstrup, S.C., Born, E.W., Calvert, W., Derocher, A.E., Garner, G.W., Messier, F., Stirling, I., Taylor, M.K., Wiig, Ø., and Strobeck, C. 1999. Genetic structure of the world's polar bear populations. *Molecular Ecology* 8: 1571-1584.
- Peacock, E., Taylor, M.K., Laake, J., and Stirling, I. 2013. Population ecology of polar bears in Davis Strait, Canada and Greenland. *Journal of Wildlife Management* 77: 463-476.
- Peacock, E., Sonsthagen, S.A., Obbard, M.E., Boltunov, A., Regehr, E.V., Ovsyanikov, N., Aars, J., Atkinson, S.N., Sage, G.K., Hope, A.G., Zeyl, E., Bachmann, L., Ehrich, D., Scribner, K.T., Amstrup, S.C., Belikov, S., Born, E., Derocher, A.E., Stirling, I., Taylor, M.K., Wiig, Ø., Paetkau, D., and Talbot, S.L. 2015. Implications of the circumpolar genetic structure of polar bears for their conservation in a rapidly warming Arctic. *Plos One* 10: e112021.
- Rode, K.D., Peacock, E., Taylor, M., Stirling, I., Born, E.W., Laidre, K.L. and Wiig, Ø. 2012. A tale of two polar bear populations: ice habitat, harvest, and body condition. *Population Ecology* 54: 3-18.
- Stirling, I., Calvert, W., and Andriashek, D. 1980. Population ecology studies of the polar bear in the area of southeastern Baffin Island. Canadian Wildlife Service Occasional Paper No. 44, 33 pp.

- Stirling, I. and Kiliaan, H.P.L. 1980. Population ecology studies of the polar bear in northern Labrador. Canadian Wildlife Service Occasional Paper No. 42, 21 pp.
- Taylor, M. and Lee, J. 1995. Distribution and abundance of Canadian polar bear populations: A management perspective. *Arctic* 48: 147-154.
- Taylor, M.K., Akeeagok, S., Andriashek, D., Barbour, W., Born, E.W., Calvert, W., Dean Cluff, H., Ferguson, S., Laake, J. Rosing-Asvid, A., Stirling, I., and Messier, F. 2001. Delineating Canadian and Greenland polar bear (*Ursus maritimus*) populations by cluster analysis of movements. *Can. J. Zool.* 79: 690-709.
- Tomaselli, M., Henri, D., Pagnirtung Hunters and Trappers Organization, Mayukalik Hunters and Trappers Organization, Akavak, N., Kanayuk, D., Kanayuk, R., Pitsiulak, P., Wong, P., Richardson, E., and Dyck, M. 2022. Nunavut Inuit Qaujimagatuqangit on the Health of the Davis Strait Polar Bear Population. Final Project Report.
- York, J., Dale, A., Mitchell, J., Nash, T., Snook, J., Felt, L., Taylor, M., and Dowsley, M. 2015. Labrador polar bear Traditional Ecological Knowledge final report. Torngat Wildlife, Plants and Fisheries Secretariat Ser. 2015/03 +118 + iv p.

Foxe Basin (FB)

Status and delineation

Based on decades of mark-recapture studies and satellite tracking of female bears in Western Hudson Bay (WH) and Southern Hudson Bay (SH), the Foxe Basin subpopulation appears to occur in Foxe Basin, northern Hudson Bay, and the western end of Hudson Strait (Taylor and Lee 1995; Sahanatien *et al.* 2015). The most recent mapping of satellite telemetry data, 2008-2009, indicates substantial overlap with the WH subpopulation and, to a lesser extent, with SH and DS (Peacock *et al.* 2010; Sahanatien *et al.* 2015) when the bears are out on the sea ice. During the ice-free season, polar bears are concentrated on Southampton Island and along the Wager Bay coast; however, significant numbers of bears are also encountered on the islands and coastal regions throughout the Foxe Basin area (Stapleton *et al.* 2015).

A 1994 subpopulation abundance estimate of 2,197 (95% CI: 1,989-2,405) was calculated (Taylor *et al.* 2006) from a mark-recapture analysis based on tetracycline biomarkers where the marking effort was conducted during the ice-free season, and distributed throughout the entire area. TEK gathered from consultations conducted in Foxe Basin communities between 2004 and 2012 suggested that the subpopulation of polar bears had increased since that initial survey. During a comprehensive summertime aerial survey in 2009 and 2010 (based on distance sampling and double-observer estimation) covering about 40,000 km each year, 816 and 1,003 bears were observed, respectively (Stapleton *et al.* 2015). This most recent study yielded an abundance estimate of 2,585 (95% CI: 2,096 – 3,189) polar bears (Stapleton *et al.* 2015), which is not statistically different from the 1994 estimate and indicated a stable population.

Habitat

Fragmentation of sea ice has increased, and total concentration and ice-floe size has decreased in FB over the last 25 years (Sahanatien and Derocher 2012), which has resulted in a reduction in sea ice habitat for polar bears (Stern and Laidre 2016). Stirling and Parkinson (2006) predicted eventual population decline based on past and predicted changes in ice habitat for polar bears but no direct evidence could be provided during the 2009-2010 aerial survey that would suggest bears of FB are currently affected negatively by climatic change (Stapleton *et al.* 2015).

Harvest management

Foxe Basin is shared between Nunavut and Québec.

Nunavut

In response to the subpopulation estimate from 1994, harvest levels in Nunavut were reduced in 1996 from 137 to 96 bears/year to permit a slow recovery of this subpopulation. After consultations in 2005, the Nunavut quota was increased to a level consistent with the increasing trend observed by Inuit and a subpopulation level estimated at 2,300 bears (106/year). The Nunavut Total Allowable Harvest (TAH) was increased from 106 to 123 bears/year in 2014/2015 based on the 2009-10 aerial survey results suggesting that the subpopulation could withstand a higher removal rate.

In September 2019, a new Nunavut Polar Bear Co-management Plan was approved by the Nunavut Wildlife Management Board, following five years of consultation and development with co-management partners in Nunavut. The Plan replaced the Memoranda of Understanding that had previously been in place. Concomitant with the approval of the new Plan, and in response to public and stakeholder feedback, Nunavut changed the sex-ratio of the harvest. Beginning with the 2019/2020 harvest season, up to 50% of a community's quota can be harvested as females, which replaces the previous 2:1 male-biased harvest; no changes to existing community TAHs were made. Therefore, there is a potential that the biological risk of negative population outcomes due to harvest will increase because more adult female polar bears could be taken and these bears are the most important contributors to population growth. The GN is monitoring how this implemented change in harvest sex-ratio will affect the sex-ratio of harvested bears with the average female proportion of harvest in 2019/2020 until 2021/2022 equaling 37.9%. Females have been underharvested relative to the annual recommended quota by approximately 11.3% on average during 2017/2018 – 2021/2022 (5 yr) time period. In February 2022, the Polar Bear Flexible Quota System was replaced by the Harvest Administration and Credit Calculation System. The up to 1:1 harvest sex-ratio is still in place.

Québec

As of 2020/2021 harvest season, there are no quotas or mandatory reporting in Quebec for FB. Voluntary reported harvest in northern Quebec averaged 3 bears/year between 2008/2009 and 2021/2022. The reporting rate by the Nunavik communities harvesting within FB is unknown but it is believed to be the lowest in Nunavik.

Indigenous Knowledge

The *Nunavut Inuit Qaujimagatuqangit on the Health of the Davis Strait Polar Bear Population* research project included interviews and focus group studies with Inuit in Kimmirut (Tomaselli *et al.* 2022). While the study focused on the Davis Strait subpopulation, Kimmirutmiut respondents discussed observations and expertise of hunting and camping areas along the southern coastal areas of Baffin Island that overlap with the Foxe Basin subpopulation. All Kimmirutmiut respondents reported that polar bears have increased in abundance since the 1940s to the 2010s. Many respondents also reported an increase in polar bear abundance since the 1980s, although two respondents reported that they have observed a recent decline in abundance since the 2000s.

Protected areas

Ukkusiksalik is a National Park in Nunavut that protects an area of bear summer concentration from development; currently there are few tourists to this National Park.

References

- Peacock, E., Derocher, A.E., Lunn, N.J., and Obbard, M.E. 2010. Polar bear ecology and management in Hudson Bay in the face of climate change. Pp. 93–115 In Ferguson, S.H., Loseto, L.L., and Mallory, M.L. (eds.). *A Little Less Arctic: Top Predators in the World's Largest Northern Inland Sea*. Springer, New York, New York, USA.
- Sahanatien, V., and Derocher, A. E. 2012. Monitoring sea ice habitat fragmentation for polar bear conservation. *Animal Conservation* **15**:397–406.
- Sahanatien, V., Peacock, E., and Derocher, A.E. 2015. Population substructure and space use of Foxe Basin polar bears. *Ecology and Evolution* **5**:2851–2864.
- Stapleton, S., Peacock, E., and Garshelis, D. 2015. Aerial surveys suggest long-term stability in the seasonally ice-free Foxe Basin (Nunavut) polar bear population. *Marine Mammal Science* **32**:181–201.
- Stern, H.L., and Laidre, K.L. 2016. Sea-ice indicators of polar bear habitat. *The Cryosphere* **10**:2027–2041.
- Stirling, I., and Parkinson, C.L. 2006. Possible effects of climate warming on selected populations of polar bears (*Ursus maritimus*) in the Canadian Arctic. *Arctic* **59**:261–275.
- Taylor, M.K., and Lee, J. 1995. Distribution and abundance of Canadian polar bear populations - a management perspective. *Arctic* **48**:147–154.
- Taylor, M.K., Lee, J., Laake, J., and McLoughlin, P.D. 2006. Estimating population size of polar bears in Foxe Basin, Nunavut using tetracycline biomarkers. File Report, Department of Environment, Government of Nunavut. Igloolik, Nunavut, Canada. 13 pp.
- Tomaselli, M., Henri, D., Pagnirtung Hunters and Trappers Organization, Mayukalik Hunters and Trappers Organization, Akavak, N., Kanayuk, D., Kanayuk, R., Pitsiulak, P., Wong, P., Richardson, E., and Dyck, M. 2022. Nunavut Inuit Qaujimagatuqangit on the Health of the Davis Strait Polar Bear Population. Final Project Report.

Gulf of Boothia (GB)

Status and delineation

The boundaries of the Gulf of Boothia subpopulation are based on genetic studies (Paetkau *et al.* 1999; Campagna *et al.* 2013; Peacock *et al.* 2015; Malenfant *et al.* 2016), movements of tagged bears (Stirling *et al.* 1978; Taylor and Lee 1995), radiotelemetry in GB and adjacent areas (Taylor *et al.* 2001), and interpretations by local Inuit hunters of how local conditions influence the movements of polar bears in the area. Genetic signatures from Gulf of Boothia bears fall within the Canadian Archipelago global genetic cluster (Peacock *et al.* 2015). An initial subpopulation estimate of 333 bears was derived from the data collected within the boundaries proposed for GB, as part of a study conducted over a larger area of the central Arctic (Furnell and Schweinsburg 1984). Although population data from this area were limited, local hunters reported that numbers remained constant or increased since the central Arctic polar bear survey in the late 1970s and early 1980s. Based on TEK, recognition of sampling deficiencies, and polar bear densities in other areas, an interim subpopulation estimate of 900 was established in the 1990s. Following the completion of a mark-recapture inventory in spring 2000, the subpopulation abundance was estimated to be $1,592 \pm 361$ bears (Taylor *et al.* 2009). Natural survival and recruitment rates were estimated at values higher than the previous standardized estimates (Taylor *et al.* 1987). Taylor *et al.* (2009) concluded that the subpopulation was increasing in 2000, as a result of high intrinsic rate of growth and low harvest. Harvest rates were increased in 2005 based on the 2000 population estimate and the population was believed to be stable.

A three-year, genetic mark-recapture population inventory study was conducted between 2015 and 2017. Results of live-capture dead-recovery models suggest a mean abundance estimate of $1,525 \pm 294$ bears for the period 2015-2017, which was similar to the previous mean abundance estimate 1998-2000 (Dyck *et al.* 2020). Mean cub-of-the-year and yearling litter sizes for the period 2015 – 2017 were 1.61 (95% confidence interval [CI] = 1.51 – 1.70) and 1.53 (95% CI = 1.41 – 1.64), respectively, with no apparent trend compared to 1998 – 2000. The mean number of yearlings per adult female for the period 2015 – 2017 was 0.36 (95% CI = 0.26 – 0.47) which suggests that GB is currently a productive polar bear subpopulation, despite sea ice change. This is consistent with the finding that polar bear body condition (i.e., fatness) in the spring improved between the periods 1998 – 2000 and 2015 – 2017. The results for subpopulation size and trend should be interpreted with caution because the genetic mark-recapture study 2015 – 2017 did not include movement data. Thus, the estimate of abundance reflects the “superpopulation” (e.g., it includes all bears that use the GB management area, some of which spend time in other subpopulations).

Harvest management

The GB subpopulation is managed solely by Nunavut. The GB quota was increased in 2005 from 40 bears to 74 bears, based on IK of increasing numbers of bears (Keith *et al.* 2005) and the results of the 1998 - 2000 study (Taylor *et al.* 2009). The quota was increased in December 2022 to 84 bears as per decision by the Nunavut Wildlife Management Board based on the results of the 2015-2017 study. Harvest of bears in GB has generally been below the recommended TAH, with an annual mean harvest of 63.6 bears since 2005.

In September 2019, the new Nunavut Polar Bear Co-management Plan was approved by the Nunavut Wildlife Management Board, following five years of consultation and development with co-management partners in Nunavut. The Plan replaced the Memoranda of Understanding that had been in place. Concomitant with the approval of the new Plan, and in response to public and stakeholder feedback, Nunavut changed the allowable harvest sex-ratio. Beginning with the 2019/2020 harvest season, up to 50% of a community's quota can be harvested as females without entering an overharvest situation. This replaces the 2:1 male-to-female harvest sex-ratio and no changes to existing community TAHs were made. There is a potential that the biological risk of negative population outcomes due to harvest will increase because adult female polar bears are the most important contributors to population growth. The GN is monitoring how this implemented change in harvest sex-ratio will affect the sex-ratio of harvested bears with the average female proportion of harvest in 2019/2020 and 2021/2022 equaling 33.8%. Females have been underharvested relative to the annual recommended quota by approximately 25.2% on average during 2017/2018 – 2021/2022 (5 yr) time-period. In February 2022, the Polar Bear Flexible Quota System was replaced by the Harvest Administration and Credit Calculation System. The up to 1:1 harvest sex-ratio is still in place.

Indigenous Knowledge

Keith *et al.* (2005) report that the number of bears in GB increased since the mid 1980's. To complement their recent scientific assessment of the Gulf of Boothia polar bear subpopulation, the Government of Nunavut Department of Environment contracted consultants (Trailmark Systems Inc.) to conduct an Inuit Knowledge study developed with Inuit communities who harvest polar bears from Gulf of Boothia. From May to August 2020, active hunters and elders were interviewed from Gjoa Haven, Taloyoak, Kugaaruk, Nauyasat, Igloodik, and Sanirajak remotely to document their knowledge of polar bear ecology, population changes (including relationships to humans), and management perspectives and considerations. Interview participants reported increasing in the number of overall bears, number of females and young bears, and number of bear encounters (Wong *et al.* 2021). Interviewees were concerned about harvest regulations that fail to take into account increasing bear numbers and human relationships to bears, from a cultural perspective. An appreciation and better inclusion of IK is needed in bear management, which will inform how decision-making impacts animals, as well as the livelihood of the communities who co-exist with them.

References

- Campagna, L., Van Coeverden de Groot, P.J., Saunders, B.L., Atkinson, S.N, Weber, D.S., Dyck, M.G., Boag, P.T., and Loughheed, S.C. 2013. Extensive sampling of polar bears (*Ursus maritimus*) in the Northwest Passage (Canadian Arctic Archipelago) reveals population differentiation across multiple spatial and temporal scales. *Ecology and Evolution* **3**:3152–3165.
- Dyck, M., Regehr, E.V., and Ware, J.V. 2020. Assessment of abundance for the Gulf of Boothia polar bear subpopulation using genetic mark-recapture. Final Report, Government of Nunavut, Department of Environment, Igloodik. 74 pp.
- Furnell, D. J. and Schweinsburg, R. E. 1984. Population-dynamics of central Canadian arctic island polar bears. *Journal of Wildlife Management* **48**:722–728.

- Keith, D., Arqvik, J., Kamookak, L. and Ameralik, J. 2005. *Inuit Qaujimaningit Nanurnut: Inuit Knowledge of Polar Bears*. Gjoa Haven Hunters and Trappers and CCI Press, Edmonton, Alberta, Canada.
- Malenfant, R.M., Davis, C.S., Cullingham, C.I., and Coltman, D.W. 2016 Circumpolar genetic structure and recent gene flow of polar bears: a reanalysis. *PLoS ONE* **11**(3):e0148967, <https://doi.org/10.1371/journal.pone.0148967>.
- Paetkau, D., Amstrup, S.C., Born, E.W., Calvert, W., Derocher, A.E., Garner, G.W., Messier, F., Stirling, I., Taylor, M.K., Wiig, Ø., and Strobeck, C. 1999. Genetic structure of the world's polar bear populations. *Molecular Ecology* **8**:1571–1584.
- Peacock, E., Sonsthagen, S.A., Obbard, M.E., Boltunov, A., Regehr, E.V., Ovsyanikov, N., Aars, J., Atkinson, S.N., Sage, G.K., Hope, A.G., Zeyl, E., Bachmann, L., Ehrich, D., Scribner, K.T., Amstrup, S.C., Belikov, S., Born, E., Derocher, A.E., Stirling, I., Taylor, M.K., Wiig, Ø., Paetkau, D., and Talbot, S.L. 2015. Implications of the circumpolar genetic structure of polar bears for their conservation in a rapidly warming Arctic. *PLoS ONE* **10**:e112021.
- Stirling, I., Schweinsburg, R.E., Calvert, W., and Killian, H.P.L. 1978. Population ecology studies of the polar bear along the proposed Arctic Islands Gas Pipeline Route, Final Report. Environmental Management Service, Department of Environment, Alberta, Canada. 93 pp.
- Taylor, M.K., and Lee, J. 1995. Distribution and abundance of Canadian polar bear populations - a management perspective. *Arctic* **48**:147–154.
- Taylor, M.K., Akeeagok, S., Andriashek, D., Barbour, W., Born, E.W., Calvert, W., Cluff, H.D., Ferguson, S., Laake, J., Rosing-Asvid, A., Stirling, I., and Messier, F. 2001. Delineating Canadian and Greenland polar bear (*Ursus maritimus*) populations by cluster analysis of movements. *Canadian Journal of Zoology* **79**:690–709.
- Taylor, M.K., DeMaster, D.P., Bunnell, F.L., and Schweinsburg, R.E. 1987. Modeling the sustainable harvest of polar bears. *Journal of Wildlife Management* **51**:811–820.
- Taylor, M.K., Laake, J., McLoughlin, P.D., Cluff, H.D., and Messier, F. 2009. Demography and population viability of polar bears in the Gulf of Boothia, Nunavut. *Marine Mammal Science* **25**:778–796.
- Wong, P.B.Y, Gjoa Haven Hunters and Trappers Organization, Spence Bay Hunters and Trappers Organization, Kurtairojuark Hunters and Trappers Organization, Aiviq Hunters and Trappers Organization, Igloodik Hunters and Trappers Organization, and Hall Beach Hunters and Trappers Organization. 2021. Inuit Qaujimajatuqangit of Gulf of Boothia Polar Bears. Final Project Report.

Kane Basin (KB)

Status and delineation

Based on the movements of adult females with satellite collars and recaptures of tagged animals, the boundaries of the Kane Basin subpopulation include the North Water Polynya to the south, the Kennedy Channel to the north and Greenland and Ellesmere Island to the east and west (Taylor *et al.* 2001). Polar bears in KB do not differ genetically from those in Baffin Bay (Paetkau *et al.* 1999; Peacock *et al.* 2015). The size of the subpopulation was estimated to be 164 ± 35 (SE) for 1994 – 1997 by Taylor *et al.* (2008). The intrinsic natural rate of growth for KB polar bears was estimated to be low at 1.009 (SE, 0.010) (Taylor *et al.* 2008), likely because of large expanses of multi-year ice and low population density of seals (Born *et al.* 2004). A genetic mark-recapture survey (via biopsy darting) and aerial survey were completed in 2014 resulting in a new population estimate, survival rates, and habitat use analyses (SWG 2016). Using genetic mark-recapture, the estimated abundance of the KB subpopulation was 357 polar bears (95% CI: 221 – 493) for 2013 – 2014. More bears were documented in the eastern regions of the KB subpopulation during 2012 – 2014 than during 1994-1997. The difference in distribution between the 1990s and 2010s may reflect differences in spatial distribution of bears, possibly influenced by reduced hunting pressure by Greenland in eastern KB but also some differences in sampling protocols. An estimate of abundance based on a springtime 2014 aerial survey in KB was 206 bears (95% lognormal CI: 83 - 510). However, due to insufficient coverage of offshore polar bear habitat, this estimate is likely negatively biased. The total number of bears marked during studies in 2012-2013 in KB was equivalent to ~25% of the estimated population size. Despite this, documented cases of emigration comprised < 4% of recaptures and recoveries in KB.

Changing sea-ice conditions have resulted in broad movement and habitat use patterns of KB bears that are more similar to those of bears in seasonal sea-ice ecoregions. The size of the subpopulation range has expanded since the 1990s in all seasons, especially in summer (June-September) where ranges doubled between the 1990s and the 2000s. Land use in KB during summer remains intermittent because some sea ice remains inside fjords and coastal areas. Reproductive metrics for KB were comparable between the 1990s and 2010s sampling periods. Body condition in KB appeared to have slightly improved between sampling periods (see SWG 2016; Laidre *et al.* 2020). Overall, the data on abundance when considered with data on movements, condition, and reproduction, suggest that the subpopulation has increased.

Harvest management

In 2009, the Canada-Greenland Joint Commission (JC) was established to manage polar bears within Kane Basin and Baffin Bay collaboratively, and to ensure their conservation. After the new population assessment concluded, a harvest risk assessment of the KB polar bear subpopulation was completed for the JC by the SWG to provide various harvest scenarios that could guide management decisions (Regehr *et al.* 2017). The harvest risk analyses incorporated various demographic approaches and vital rates based on polar bear life history, and potential effects of future sea-ice conditions on polar bear population size and status through projected trends in carrying capacity. The SWG harvest risk analysis recommended a possible removal of up to 10 (2.8%) bears from the KB polar bear subpopulation, which resulted from the high uncertainty in vital rates and overall small sample sizes during the population study. The current quota for KB is 9 bears for Greenland, and 5 for Nunavut per year. So far, the combined catches from Nunavut and Greenland for the KB subpopulation have remained well below 10 polar bears per year.

In September 2019, the new Nunavut Polar Bear Co-management Plan was approved by the Nunavut Wildlife Management Board. With the approval of the new plan, a change in the existing Nunavut polar bear harvest management system occurred: beginning with the 2019/2020 harvest season up to 50% of the communities' quota can be harvested as females without entering into an overharvest situation. The 2:1 male-to-female harvest sex ratio was changed to a 'up to 1:1 male-to-female' harvest by increasing the number of females that can be harvested (i.e., maintaining overall harvest level). There is a potential that the biological risk of negative population outcomes due to harvest will increase because adult female polar bears are the most important contributors to population growth. How this administrative change in harvest sex-ratio is affecting the sex ratio of harvested bears is being monitored by the Government of Nunavut. Harvest in Nunavut has been zero bears for the last 5 years. In February 2022, the Polar Bear Flexible Quota System was replaced by the Harvest Administration and Credit Calculation System. The up to 1:1 harvest sex-ratio is still in place.

References

- Born, E.W., Teilmann, J., Acquarone, M., and Rig  t, F.F. 2004. Habitat use of ringed seals (*Phoca hispida*) in the North Water area (North Baffin Bay). *Arctic* **57**:129–142.
- Laidre, K.L., Atkinson, S.N., Regehr, E. V, Stern, H.L., Born, E.W., Wiig,   ., Lunn, N.J., Dyck, M., Heagerty, P. and Cohen, B.R. 2020. Transient benefits of climate change for a high-Arctic polar bear (*Ursus maritimus*) subpopulation. *Global Change Biology* **26**:6251–6265.
- Paetkau, D., Amstrup, S.C., Born, E.W., Calvert, W., Derocher, A.E., Garner, G.W., Messier, F., Stirling, I., Taylor, M.K., Wiig,   ., and Strobeck, C. 1999. Genetic structure of the world's polar bear populations. *Molecular Ecology* **8**:1571–1584.
- Peacock, E., Sonsthagen, S.A., Obbard, M.E., Boltunov, A., Regehr, E.V., Ovsyanikov, N., Aars, J., Atkinson, S.N., Sage, G.K., Hope, A.G., Zeyl, E., Bachmann, L., Ehrich, D., Scribner, K.T., Amstrup, S.C., Belikov, S., Born, E., Derocher, A.E., Stirling, I., Taylor, M.K., Wiig,   ., Paetkau, D., and Talbot, S.L. 2015. Implications of the circumpolar genetic structure of polar bears for their conservation in a rapidly warming Arctic. *Plos One* **10**:e112021.

Lancaster Sound (LS)

Status and delineation

Information on the movements of adult female polar bears monitored by satellite radio-collars, and mark-recapture data has shown that this subpopulation is distinct from the adjoining Viscount Melville Sound, M'Clintock Channel, Gulf of Boothia, Baffin Bay, and Norwegian Bay subpopulations (Taylor *et al.* 2001). This distinction is supported by genetic data (Paetkau *et al.* 1999; Malenfant *et al.* 2016; Peacock *et al.* 2015); polar bears in LS belong to the *Canadian Archipelago* genetic cluster (Malenfant *et al.* 2016; Peacock *et al.* 2015). Survival rates of the pooled Norwegian Bay and LS populations were used in a population viability analysis (PVA) to minimize sampling errors; the available subpopulation estimate of $2,541 \pm 391$ is based on an analysis of both historical and current mark-recapture data to 1997 (Taylor *et al.* 2008). Taylor *et al.* (2008) estimated survival and recruitment parameters that suggest this subpopulation has a lower renewal rate than previously estimated. However, what effect this may, or may not, have on the present population demographics is not known, especially under changing sea-ice conditions. Currently, the population data are outdated but the population is thought to be stable based on local traditional knowledge (Canadian Wildlife Service 2009). A new population study was initiated in 2021 and is ongoing, with the survey component having been completed spring 2023.

Harvest management

Lancaster Sound is managed solely by Nunavut. The polar bear harvest in LS has been more male-selective for decades as compared to any other Nunavut subpopulations, in part because of guided sport hunts. Sport hunting has been an important economic activity for hamlets in Nunavut, but more so for communities that harvest from LS – approximately 40% of all Nunavut sport hunts occur in LS (Government of Nunavut, unpublished data). With the 2008 ban on hide importation into the US, sport-hunting tourism declined from around 45% of the LS harvest to approximately 16% of the LS harvest. Over time, sport hunting has increased again and between 2016/2017 – 2018/2019, sport hunts were approximately 31% of the LS harvest. COVID-19 reduced sport hunts to essentially zero for LS in 2019/2020 and 2020/2021. The mean male proportion of the harvest since 1983 has been about 0.72, which is higher than the prescribed 2:1 harvest sex ratio (Government of Nunavut, unpublished data). McLoughlin *et al.* (2005) examined the conservation risk of male-selective harvesting of polar bears but there are no sufficient data to examine the long-term effects of that harvest regimen on LS polar bear demography. However, data collected during the current subpopulation study may aid such endeavors. The quota increased in 2004/2005 from 78 to 85 bears annually; the mean removal per year from 2004/2005 to 2021/2022 has been 80.8 bears (range: 70 – 94 bears).

In September 2019, the new Nunavut Polar Bear Co-management Plan was approved by the Nunavut Wildlife Management Board, following five years of consultation and development with co-management partners in Nunavut. The Plan replaced the Memoranda of Understanding that had been in place. Concomitant with the approval of the new Plan, and in response to public and stakeholder feedback, Nunavut changed the allowable harvest sex-ratio. Beginning with the 2019/2020 harvest season, up to 50% of a community's quota can be harvested as females without entering an overharvest situation. This replaces the 2:1 male-to-female harvest sex-ratio and no changes to existing community TAHs were made. There is a potential that the biological risk of negative population outcomes due to harvest will increase because adult female polar bears are the most important contributors to population growth. The GN is monitoring how this implemented change in harvest sex-ratio will affect the sex-ratio of harvested bears. The average female proportion of harvest in 2019/2020 until 2021/2022 is 24.7%. Females have been underharvested relative to the annual recommended quota by approximately 42.1% on average during 2017/2018 – 2021/2022 (5 yr) time period. In February 2022, the Polar Bear Flexible Quota System was replaced by the Harvest Administration and Credit Calculation System. The up to 1:1 harvest sex-ratio is still in place.

Protected areas

Qausuittuq National Park and portions of Sirmilik National Park provide protection of denning and summering polar bear habitat. One of Canada's largest National Marine Conservation Areas was created in Lancaster Sound in 2019; *Tallrutiup Imanga* covers large areas of central and eastern LS.

References

- Canadian Wildlife Service. 2009. Nunavut consultation report – Consultations on the proposed listing of the Polar Bear as Special Concern under the Species at Risk Act. Report submitted to the Nunavut Wildlife Management Board in accordance with Step 3.8 of the Memorandum of Understanding to Harmonize the Designation of Endangered Species under the Nunavut Land Claims Agreement and the Species at Risk Act, 249 pp. [available at: <http://assembly.nu.ca/library/Edocs/2009/001149-e.pdf>].
- Malenfant, R.M., Davis, C.S., Cullingham, C.I., and Coltman, D.W. 2016 Circumpolar genetic structure and recent gene flow of polar bears: a reanalysis. *PLoS ONE* **11**(3):e0148967, <https://doi.org/10.1371/journal.pone.0148967>.
- McLoughlin, P.D., Taylor, M.K., and Messier, F. 2005. Conservation risks of male-selective harvest for mammals with low reproductive potential. *Journal of Wildlife Management* **69**:1592–1600.
- Paetkau, D., Amstrup, S.C., Born, E.W., Calvert, W., Derocher, A.E., Garner, G.W., Messier, F., Stirling, I., Taylor, M.K., Wiig, Ø., and Strobeck, C. 1999. Genetic structure of the world's polar bear populations. *Molecular Ecology* **8**:1571–1584.
- Peacock, E., Sonsthagen, S.A., Obbard, M.E., Boltunov, A., Regehr, E.V., Ovsyanikov, N., Aars, J., Atkinson, S.N., Sage, G.K., Hope, A.G., Zeyl, E., Bachmann, L., Ehrich, D., Scribner, K.T., Amstrup, S.C., Belikov, S., Born, E., Derocher, A.E., Stirling, I., Taylor, M.K., Wiig, Ø., Paetkau, D., and Talbot, S.L. 2015. Implications of the circumpolar genetic structure of polar bears for their conservation in a rapidly warming Arctic. *PLoS ONE* **10**:e112021.
- Stirling, I., Calvert, W., and Andriashek, D. 1984. Polar bear ecology and environmental considerations in the Canadian High Arctic. Pp. 201-222 In Olson, R., Geddes F. and Hastings, R. (eds.). *Northern Ecology and Resource Management*. University of Alberta Press, Edmonton, Canada.

- Taylor, M.K., Akeagok, S., Andriashek, D., Barbour, W., Born, E.W., Calvert, W., Cluff, H.D., Ferguson, S., Laake, J., Rosing-Asvid, A., Stirling, I., and Messier, F. 2001. Delineating Canadian and Greenland polar bear (*Ursus maritimus*) populations by cluster analysis of movements. *Canadian Journal of Zoology* **79**:690–709.
- Taylor, M.K., Laake, J., McLoughlin, P.D., Cluff, H.D., and Messier, F. 2008. Mark-recapture and stochastic population models for polar bears of the high Arctic. *Arctic* **61**:143–152.

M'Clintock Channel (MC)

Status and delineation

The current population boundaries for the M'Clintock Channel (MC) subpopulation are based on recovery of tagged bears, movements of adult females with satellite radio-collars in adjacent areas (Taylor and Lee 1995, Taylor *et al.* 2001), and genetics (Paetkau *et al.* 1999; Campagna *et al.* 2013; Peacock *et al.* 2015, Malenfant *et al.* 2016). An initial physical mark-recapture study was carried out from 1973 – 1978 (Furnell and Schweinsburg 1984) in areas that included portions of what is now known as MC and the adjacent Gulf of Boothia subpopulation together. The total abundance estimated for the area was 1081 bears, but the estimate was known to be biased by non-representative sampling. It was subsequently increased to 900 for GB and 900 for MC based on local indigenous knowledge and back-calculations to determine abundance levels necessary to sustain the existing subsistence harvest levels (Aars *et al.* 2006; Taylor *et al.* 2006). In the mid-1990s, the MC estimate was revised downwards to 700 based on hunter reports of reduced densities of polar bears (Aars *et al.* 2006; Taylor *et al.* 2006). Following the completion of a mark-recapture inventory in spring 1998-2000, which covered the currently delineated subpopulation area, the abundance was estimated at 284 ± 59.3 bears (Taylor *et al.* 2006). Natural survival and recruitment rates were estimated at values lower than previous standardized estimates (Taylor *et al.* 1987).

A new 3-year genetic mark-recapture study was conducted between 2014 and 2016, indicating that the population increased to 716 bears (95% Credible Interval: 545-955; Dyck *et al.* 2020). Body condition of bears between 1998-2000 and 2014-2016 improved indicating that temporal and spatial changes in sea ice conditions likely improved marine productivity and benefitted bears. The results for subpopulation size and trend should be interpreted with caution because the 2014 – 2016 genetic mark-recapture study did not include movement data (radio-collars) and thus, the estimate of abundance reflects the “superpopulation” (e.g., it includes all bears that use the MC management area, some of which spend time in other subpopulations).

Habitat

Similar to habitat in GB, Barber and Iacozza (2004) found no trends in ringed seal habitat or sea ice conditions from 1980 to 2000 for the MC area. A general trend has been detected for earlier break-up and delayed freeze-up (Stern and Laidre 2016; Markus *et al.* 2009), with multi-year ice shifting to lesser occurrence than annual ice as compared to previous decades (e.g. Howell *et al.* 2015). Habitat quality could be improved over the short-term as multi-year ice declines.

There is the potential for increased shipping activities in parts of the MC area as the Northwest Passage becomes more navigable in the future.

Harvest management

The MC subpopulation is managed solely by Nunavut. Past harvests in MC of 34 bears per year between 1979 – 1999 were considered unsustainable (Aars *et al.* 2006; Taylor *et al.* 2006) based on local knowledge of decreased bear density and the results of the 1998 – 2000 capture mark-recapture study. Consequently, a harvest moratorium was put in place for the 2001/2002 and 2002/2003 harvest seasons followed by a quota of 3 bears annually until 2015. Local hunters observed more bears during the late 2000's, which resulted in an increase of the annual quota from 3 to 12 bears in 2015/2016 at a 2:1 male-biased harvest. In 2021, the TAH was increased to 21 bears based on the updated population estimate and indigenous knowledge.

In September 2019, the new Nunavut Polar Bear Co-management Plan was approved by the Nunavut Wildlife Management Board, following five years of consultation and development with co-management partners in Nunavut. The Plan replaced the Memoranda of Understanding that had been in place. Concomitant with the approval of the new Plan, and in response to public and stakeholder feedback, Nunavut changed the allowable harvest sex-ratio. Beginning with the 2019/2020 harvest season, up to 50% of a community's quota can be harvested as females without entering an overharvest situation. This replaces the 2:1 male-to-female harvest sex-ratio and no changes to existing community TAHs were made. There is a potential that the biological risk of negative population outcomes due to harvest will increase because adult female polar bears are the most important contributors to population growth. The effects from the change in harvest sex ratio on the actual sex ratio of harvested bears will continue to be monitored by the Government of Nunavut. The average female proportion of harvest in 2019/2020 until 2021/2022 is 19%. Females have been underharvested relative to the annual recommended quota by approximately 35% on average during 2017/2018 – 2021/2022 (5 yr) time period. In February 2022, the Polar Bear Flexible Quota System was replaced by the Harvest Administration and Credit Calculation System. The up to 1:1 harvest sex-ratio is still in place.

Indigenous Knowledge

The Government of Nunavut contracted Trailmark Systems Inc. consultants to conduct an Inuit Knowledge study led by communities who harvest polar bears from MC subpopulation. In May and June 2020, active hunters and elders were interviewed from Cambridge Bay, Gjoa Haven, and Taloyoak remotely to document their knowledge of polar bear ecology, population changes (including human-animal relationships), and management perspectives and recommendations. Interviewees voiced concerns over changing human-bear relationships that have led to more aggressive bears and increasing bear numbers in M'Clintock Channel, which—combined with too few hunting tags—pose a threat to human safety (Wong *et al.* 2021). Interviewees also consider the inclusion of Inuit perspectives and traditions in research and management inadequate to-date. Decision-makers and researchers need to improve their understanding of Inuit knowledge to fully consider and include Inuit Knowledge in bear research and management. These efforts will encourage more balanced, culturally appropriate, and sustainable management practices that are supported by community members.

References

Aars, J., Lunn, N.J., and Derocher, A.E. (eds.). 2006. *Polar Bears: Proceedings of the 14th Working Meeting of the IUCN/SSC Polar Bear Specialist Group, 20-24 June 2005, Seattle, Washington*. Gland, Switzerland and Cambridge, UK.

- Barber, D.G. and Iacozza, J. 2004. Historical analysis of sea ice conditions in M'Clintock channel and the Gulf of Boothia, Nunavut: Implications for ringed seal and polar bear habitat. *Arctic* **57**:1–14.
- Campagna, L., Van Coeverden de Groot, P.J., Saunders, B.L., Atkinson, S.N., Weber, D.S., Dyck, M.G., Boag, P.T., and Lougheed, S.C. 2013. Extensive sampling of polar bears (*Ursus maritimus*) in the Northwest Passage (Canadian Arctic Archipelago) reveals population differentiation across multiple spatial and temporal scales. *Ecology and Evolution* **3**:3152–3165.
- Dyck, M., Lukacs, P., and Ware, J.V. 2020. Re-estimating the abundance of a recovering polar bear subpopulation by genetic mark-recapture in M'Clintock Channel, Nunavut, Canada. Final Report, Government of Nunavut, Department of Environment, Iglulik, 79 pp.
- Furnell, D. J. and Schweinsburg, R.E. 1984. Population-dynamics of central Canadian arctic island polar bears. *Journal of Wildlife Management* **48**:722–728.
- Howell, S.E.L., Derksen, C., Pizzolato, L., and Brady, M. 2015. Multiyear ice replenishment in the Canadian Arctic Archipelago: 1997–2013. *Journal of Geophysical Research: Oceans* **120**:1623–1637.
- Malenfant, R.M., Davis, C.S., Cullingham, C.I., and Coltman, D.W. 2016. Circumpolar genetic structure and recent gene flow of polar bears: a reanalysis. *PLoS ONE* **11**(3):e0148967, <https://doi.org/10.1371/journal.pone.0148967>.
- Markus, T., Stroeve, J.C., and Miller, J. 2009. Recent changes in Arctic sea ice melt onset, freezeup, and melt season length. *Journal of Geophysical Research* **114**:C12024, <https://doi.org/10.1029/2009JC005436>.
- Paetkau, D., Amstrup, S.C., Born, E.W., Calvert, W., Derocher, A.E., Garner, G.W., Messier, F., Stirling, I., Taylor, M.K., Wiig, Ø., and Strobeck, C. 1999. Genetic structure of the world's polar bear populations. *Molecular Ecology* **8**:1571–1584.
- Peacock, E., Sonsthagen, S.A., Obbard, M.E., Boltunov, A., Regehr, E.V., Ovsyanikov, N., Aars, J., Atkinson, S.N., Sage, G.K., Hope, A.G., Zeyl, E., Bachmann, L., Ehrich, D., Scribner, K.T., Amstrup, S.C., Belikov, S., Born, E., Derocher, A.E., Stirling, I., Taylor, M.K., Wiig, Ø., Paetkau, D., and Talbot, S.L. 2015. Implications of the circumpolar genetic structure of polar bears for their conservation in a rapidly warming Arctic. *PLoS ONE* **10**:e112021.
- Stern, H.L., and Laidre, K.L. 2016. Sea-ice indicators of polar bear habitat. *The Cryosphere* **10**:2027–2041.
- Taylor, M.K., and Lee, J. 1995. Distribution and abundance of Canadian polar bear populations - a management perspective. *Arctic* **48**:147–154.
- Taylor, M.K., Akeeagok, S., Andriashek, D., Barbour, W., Born, E.W., Calvert, W., Cluff, H.D., Ferguson, S., Laake, J., Rosing-Asvid, A., Stirling, I., and Messier, F. 2001. Delineating Canadian and Greenland polar bear (*Ursus maritimus*) populations by cluster analysis of movements. *Canadian Journal of Zoology* **79**:690–709.
- Taylor, M.K., DeMaster, D.P., Bunnell, F.L., and Schweinsburg, R.E. 1987. Modeling the sustainable harvest of polar bears. *Journal of Wildlife Management* **51**:811–820.
- Taylor, M.K., Laake, J., McLoughlin, P.D., Cluff, H.D., and Messier, F. 2006. Demographic parameters and harvest-explicit population viability analysis for polar bears in M'Clintock Channel, Nunavut, Canada. *Journal of Wildlife Management* **70**:1667–1673.
- Wong, P.B.Y., Ekaluktutiak Hunters and Trappers Organization, Gjoa Haven Hunters and Trappers Organization, and Spence Bay Hunters and Trappers Organization. 2021. Inuit Qaujimagatuqangit of M'Clintock Channel Polar Bears. Final Project Report.

Northern Beaufort Sea (NB)

Boundary

The Northern Beaufort Sea (NB) subpopulation extends from Tuktoyaktuk (133° W) east through Amundsen Gulf and Dolphin and Union Strait to include Coronation Gulf. It covers nearly all of the Northern Beaufort Sea and into M'Clure Strait. This boundary was formally accepted by management authorities for the Northern and Southern Beaufort Sea polar bear subpopulations in 2013. The NB subpopulation includes portions of the Northwest Territories and Nunavut. The previous boundary between the Southern Beaufort (SB) and NB polar bear subpopulation existed at approximately 125°W longitude, near Pearce Point, NWT (Brower *et al.* 2002). The boundary change was proposed by researchers, resulting from radio telemetry studies that suggest this boundary did not reflect the space use patterns of bears in the eastern portion of the southern Beaufort (SB) Sea - records indicate that approximately 90% of the bears harvested near Baillie Island were actually NB bears (Amstrup *et al.* 2005). In consideration of the apparent misallocation of NB bears to the SB harvest, the Wildlife Management Advisory Council (NWT) (WMAC NWT) and Inuvialuit Game Council (IGC) consulted with communities regarding the potential to change the SB/NB boundary. As a result, in 2013/14, the boundary was moved west to 133°W longitude, near the community of Tuktoyaktuk, NWT. The proportional representation of NB versus SB bears reduces to approximately 50:50 at this longitude, thus allowing harvest to be more accurately allocated between the subpopulations. A re-analysis of 2001-2006 capture data (Regehr *et al.* 2007, Stirling *et al.* 2007) was undertaken to estimate the SB and NB subpopulations under the new boundary (Griswold *et al.* 2017). The mean number of bears moved from the SB to NB is 311, which is being used until another subpopulation estimate is available (Griswold *et al.* 2017).

Overview of management partners and management objectives

The management partners and collaborating agencies for the NB subpopulation on the Inuvialuit Settlement Region (ISR) side are the Government of the Northwest Territories, the WMAC (NWT), the IGC, and Environment and Climate Change Canada. In Nunavut, management partners include Government of Nunavut, Kugluktuk Hunters and Trappers Organization, Kitikmeot Regional Wildlife Board and the Nunavut Wildlife Management Board. Management objectives and guiding principles for the NB are outlined in the *Polar Bear Management Agreement for the North[ern] Beaufort Sea and Viscount Melville Sound Polar Bear Populations between Inuit of the Kitikmeot West Region in Nunavut and the Inuvialuit* (2006). The primary objectives of this agreement are:

- To maintain the Northern Beaufort Sea and Viscount Melville Sound polar bear populations at healthy viable levels in perpetuity, and
- To manage polar bears on a sustained yield basis in accordance with all the best information available

Under the *Species at Risk (NWT) Act*, polar bears are listed as a species of Special Concern. The *Inuvialuit Settlement Region Polar Bear Joint Management Plan* was published in 2017 for the species; the goal of this plan is to ensure the long-term persistence of healthy polar bears in the ISR while maintaining traditional Inuvialuit use.

Indigenous Knowledge

To date, a number of Indigenous knowledge (IK) studies have been completed that cover the ISR polar bear populations; the largest in scope and most recent is the 2015 book *Inuvialuit and Nanuq: A polar bear traditional knowledge study* (Joint Secretariat, 2015). It is important to note, however, there can be significant IK presented orally during events like project consultation meetings or public hearings that is often not adequately captured in a way to use as reference material. *Inuvialuit and Nanuq* (Joint Secretariat, 2015) describes Indigenous knowledge:

“The most important aspects of Indigenous knowledge concerning polar bears are the intergenerational knowledge (acquired from parents, grandparents and other elders) combined with direct experience. In general, this is what Inuvialuit mean by Traditional Knowledge (TK): personal knowledge acquired by travelling across ice, hunting seals and polar bears, running dog teams, reading wind directions, snow and cloud patterns, geographic features, currents and stars, and by intergenerational transmission.” – JS 2015, p. 9

Inuvialuit note that “ice conditions, the effects of climate change and polar bear behaviour are extremely complex.” (JS 2015, p. 197) Inuvialuit are reluctant to speculate about the future and long-term polar bear survival trends, given the high level of uncertainty in ecological conditions and how both bears and Inuvialuit will respond to these changes (JS 2015). Inuvialuit knowledge provides relative observations that can, in some cases, be used comparatively, to assess trends over time, or to draw a fuller picture of the NB polar bear subpopulation.

Abundance

During the verification workshop for *Inuvialuit and Nanuq* (JS 2015), consensus statements on changes to polar bear abundance over the lifetime of the TK holders were generated for each community:

- Sachs Harbour — “I don’t see the numbers going down. We’re seeing more around town, but that doesn’t mean there’s a decline in the numbers” (JS 2015, p. 184)
- Ulukhaktok — “maybe a little change, but overall about the same. Polar bear movements are always different every year. To me it’s the same, but a little bit change since when I was younger” (JS 2015, p. 184)
- Paulatuk — “The big picture is that they’re stable” – (JS 2015, p. 184)

The 2021 *Species Status Report for Polar Bear (Ursus maritimus) in the Northwest Territories* (SARC, 2021) provides a summary of other sources of Inuvialuit and local knowledge of relative polar bear and seal abundance in the Inuvialuit Settlement Region (see that SARC 2021 for further detail). This historical perspective demonstrates the complexities of polar bear ecology and that the abundance and distribution of polar bears and their prey has always been variable.

Body Condition

Inuvialuit knowledge holders in *Inuvialuit and Nanuq* (JS 2015) agreed that polar bear body condition has remained generally stable over time, despite considerable variability within and between years. Inuvialuit also indicated that, since the 1980s, there have been less really big bears observed, and the big bears aren't as fat (JS 2015).

Distribution

In some parts of the ISR, Inuvialuit knowledge holders have observed dens in different places than they were before and females with cubs have been observed entering and leaving dens at different times – this is attributed to changing weather patterns (JS 2015). Sachs Harbour knowledge holders have observed far less landfast ice than there was before the mid-1980s, which results in polar bears staying closer to the community (JS 2015). Polar bears are being observed closer to the mainland in and around Tuktoyaktuk and Paulatuk. Inuvialuit have not linked these changes in distribution to changes in polar bear abundance (JS 2015).

Climate Change

Inuvialuit see and experience climate change firsthand, noting changes to temperature, freeze-up, break-up, ice conditions, wind and storm patterns; the book *Inuvialuit and Nanuq* (JS 2015) explores Inuvialuit observations of climate change since the 1980s. Despite observations of the climate change and a nuanced understanding of polar bear ecology, Inuvialuit have not yet seen changes to polar bear abundance or condition (JS 2015). Inuvialuit consensus is that:

“For the Inuvialuit, the future cannot be predicted; it could be good or bad as far as polar bears are concerned. However, the consensus among the workshop participants was that polar bears are highly intelligent animals that can adapt to climate change because they have been adapting to many things for thousands of years.” (JS 2015, p. 196)

Scientific Knowledge

Abundance

Scientific study of the NB polar bear subpopulation extends back to the 1970s. It is worth noting that scientific polar bear research is very expensive and takes place in remote, dangerous areas where weather patterns are highly unpredictable. These factors can significantly affect study success and the frequency of population inventories.

There have been multiple population assessments conducted in the NB, and all were based upon the former subpopulation boundaries. Inventory periods and resultant population estimates during each decade are as follows (as documented in Stirling et al. (2007) except final 2006 estimate):

Table 5. Population estimates for Northern Beaufort Sea subpopulation

Inventory period	Population Estimate	99% Confidence Interval	Estimate for Management Purposes	Comments and Reference
1972-1975	745	± 246	1,200	Stirling <i>et al.</i> 1975, Estimate from Stirling et al 2007
1985-1987	847	± 141	1,200	Stirling <i>et al.</i> 1988, Estimate from Stirling et al 2007
1992-1994	289	± 62	1,200	Only area north of Norway Island covered consistently. (Lunn <i>et al.</i> 1995.) Estimate from Stirling et al 2007
2004-2006	980	± 155	1,400	Estimate from Stirling et al 2007 and Stirling et al 2011. Increase in estimate based on negative bias due to lack of capture effort in north and east portions of study area. Stirling <i>et al.</i> 2011 state estimate 1200-1300 in 2004 - 2005 more reasonable.
2006	1,291		1,711	Boundary change moves estimated 311 bears based on analysis in 2009 (Griswold et al. 2017) and estimate used for management purposes adjusted for bias in sampling.

Stirling *et al.* (2007) indicate that the estimate of bears during the 1990s was relatively quite low; however, capture effort for this period differed from other periods, and was focused in the northern portion of the subpopulation (northwest corner of Banks Island and Prince Patrick Island); the estimate was adjusted for management purposes.

The NB population estimate under the current boundary is 1,291, a number derived from the 2000s estimate with the addition of 311 bears (following analysis in 2009) that estimated the number of bears that would shift between subpopulations under the boundary change (Griswold et al. 2017).

Stirling *et al.* (2011) recognized that the estimate from the 2000s (980) was likely biased low (possibly related to changes in distribution), and suggested the population estimates of 1200-1300 in 2004 and 2005 may more accurately reflect the current number of bears in the population. Furthermore, they recognized that limited sampling in the northern portion of the study area may have led to estimates that are biased low.

The NB population estimate used for management purposes has historically and continues to be adjusted to reflect negative bias. The current estimate used for management purposes of the NB is 1,711 (WMAC (NWT) 25 July 2011).

A genetic mark-recapture survey of the Southern and Northern Beaufort Sea subpopulations is underway. The first year of field work was 2019, with three to four years of fieldwork planned in total. Field work in 2020 was postponed due to the Covid-19 pandemic, but resumed in 2021. Discussions are ongoing about the possible inclusion of additional marks for this survey using fecal samples and eDNA from polar bear tracks. In 2020, the Joint Secretariat hired contractors to develop a potential framework for incorporating IK into an Integrated Population Model (IPM) for SB and NB polar bears, with a focus on incorporating IK as a source of information.

Habitat/Climate Change

Multiple indicators of climate change impacts on sea ice have been noted for the NB polar bear subpopulation. From 1979 to 2014, researchers have observed: a declining number of ice-covered days, a declining rate of June to October sea ice concentration, and an increasing length of the summer season (Stern and Laidre 2016). The length of the summer season increased by 9.0 days from 1979 to 2014 for the NB (Stern and Laidre 2016).

Based on fatty acid signature, the diet of polar bears in the Beaufort was dominated by ringed seal, but there was variation. Most variation in bear diet was explained by longitude, reflecting spatial variation in prey availability. Sea ice conditions (extent, thickness, and seasonal duration) declined throughout the study period, and date of sea ice break-up in the preceding spring was positively correlated with female body condition and consumption of beluga whale (Florko et al. 2020).

Harvest Management

Within the ISR, harvest is carefully managed. All human-caused mortality including hunting, defense of life and property kills, industry-related mortalities and illegal kills are tracked and counted under a quota. There is mandatory reporting and submission of proof of sex and age that is enforceable under the Wildlife Act. A key aspect that ensures human caused mortality remains below TAH is a highly adaptive management system whereby information related to population abundance and trend is evaluated annually by the WMACs (NWT and NS) and the IGC and changes are recommended to HTC's or the appropriate Minister (s) when required.

Hunting in the NB has historically been focused in the Amundsen Gulf and western coast of Banks Island (with a focus near Sachs Harbour) (Usher 1976, JS 2015, GNWT unpublished data).

Within Canada, quotas were first established in NWT by the 33rd Session of the Territorial Council at Resolute Bay. The quotas were to become effective on July 1 for the 1967-68 hunting season. In the absence of data, quotas for each settlement were established by averaging the harvest of the previous 3 years and then reducing that number by a modest amount (Brower et al 2002).

The first quota increases based on scientific information were made in 1978-79 after completion of the first population study of polar bears in the Western Arctic (Stirling 1975).

Currently a combined (NU and NWT) total allowable harvest for NB is 77 bears per year, but the actual average harvest over the last five years is approximately 31 (GNWT/GN unpublished data). In NU (quota of 6/year), harvest has declined, in part because of increasing difficulty of Kugluktuk residents to reach areas where there are bears, because of changing ice conditions. Harvest of the NB subpopulation has been below the quota for several years. Changing sea ice, distance needed to travel, challenging ice conditions and cost of travel to access bears are all cited as reasons (Larry Carpenter pers. comm. 2020). Changing sea ice conditions has made it difficult for Inuvialuit to rely on established IK for planning harvest activities (JS, 2015).

Protected areas

Some denning habitat is protected in Aulavik National Park on the northern coast of Banks Island, Northwest Territories but most known maternity denning in NB occurs along the southern and western coastlines of Banks Island, and associated small offshore islands. Bears in dens are protected by Hunters and Trappers Committee by-laws and regulations.

References

- Amstrup, S. C., G. M. Durner, I. Stirling, and T. L. McDonald. 2005. Allocating harvests among polar bear stocks in the Beaufort Sea. *Arctic* 58(3):247-259.
- Brower, C.D., A. Carpenter, M.L. Branigan, W. Calvert, T. Evans, A.S. Fischbach, J.A. Nagy, S. Schliebe, and I. Stirling. 2002. The Polar Bear Management Agreement for the Southern Beaufort Sea: An evaluation of the first ten years of a unique conservation agreement. *Arctic* 55:362–372.
- Carpenter, L. 2020. Personal communication at the Polar Bear Technical Committee meeting, February 3-6, 2020.
- GNWT. 2019. Unpublished harvest data collected under Hunters and Trappers Committee By-laws and regulations.
- Griswold J., T. McDonald, M. Branigan, E. V. Regehr, and S. C. Amstrup. 2017. Southern and northern Beaufort Sea polar bear population estimates under a proposed boundary shift. Manuscript Report 265.
- Joint Secretariat. 2015. Inuvialuit and Nanuq: A Polar Bear Traditional Knowledge Study. Joint Secretariat, Inuvialuit Settlement Region. xx + 304 pp. Available online: http://www.wmacns.ca/pdfs/394_polar-bear-tk-report-low-res.pdf.
- Joint Secretariat. 2017. Inuvialuit Settlement Region Polar Bear Joint Management Plan. Joint Secretariat, Inuvialuit Settlement Region. vii + 66 pp. https://www.nwt-species-at-risk.ca/sites/default/files/isr_polar_bear_joint_management_plan_2017_final.pdf

- Lunn, N.J., I. Stirling, and D. Andriashek. 1995. Movements and Distribution of Polar Bears in the northeastern Beaufort Sea and western McClure Strait. Final report to Wildlife Management Advisory Committee. 65 p.
- Regehr, E. V., S. C. Amstrup, and I. Stirling. 2007. Polar Bear Population Status in the Southern Beaufort Sea. U.S. Geological Survey Open-File Report 2006-1337. 20 pp.
- Species at Risk Committee. 2012. Species Status Report for Polar Bear (*Ursus maritimus*) in the Northwest Territories. Species at Risk Committee, Yellowknife, NT.
- Stern, H.L., and Laidre, K.L. 2016. Sea-ice indicators of polar bear habitat. *Cryosphere*, 10: 2027–2041. doi: 10.5194/tc-10-2027-2016.
- Stirling, I., D. Andriashek, P. Latour, and W. Calvert. 1975. The distribution and abundance of polar bears in the eastern Beaufort Sea. Final Report to the Beaufort Sea Project. Fisheries and Marine Service, Department of Environment, Victoria, B.C. 59 pp.
- Stirling, I., D. Andriashek, C. Spencer and A. Derocher. 1988. Assessment of the polar bear population in the eastern Beaufort Sea. Final report to the Northern Oil and Gas Action Program. Canadian Wildlife Service. Edmonton.
- Stirling, I., T.L. McDonald, E.S. Richardson and E.V. Regehr. 2007. Polar bear population status in the Northern Beaufort Sea. USGS Administrative Report
- Stirling, I., T. L. McDonald, E. S. Richardson, E. V. Regehr, and S. C. Amstrup. 2011. Polar bear population status in the northern Beaufort Sea, Canada, 1971-2006. *Ecological Applications* 21:859-876.
- Usher, P. 1976. Inuit Land Use in the Western Canadian Arctic. Pp. 21-31 in: Inuit Land Use and Occupancy Report, Vol. 1. M. M. R. Freeman, ed. Department of Indian Affairs and Northern Development, Ottawa, ON.
- WMAC (NWT). 25 July 2011. Letter to ENR Minister re: Recommendations for northern Beaufort Sea polar bear population boundary change and Total Allowable Harvest. Inuvik, NT.

Norwegian Bay (NW)

Status and delineation

The Norwegian Bay subpopulation is bounded by heavy multi-year ice to the west, islands to the north, east, and west, and polynyas to the south (Stirling *et al.* 1993; Stirling 1997; Taylor *et al.* 2008). Data collected during mark-recapture studies, and from satellite radio-tracking of adult female polar bears, suggest that most of the polar bears in this subpopulation are concentrated along the coastal tide cracks and ridges along the north, east, and southern boundaries (Taylor *et al.* 2001). The most current (1993 – 97) estimate is 203 ± 44 (SE; Taylor *et al.* 2008). Survival rate estimates for the NW subpopulation were derived from pooled Lancaster Sound and NW data because the subpopulations are adjacent and the number of bears captured in NW was too small to generate reliable survival estimates. The NW subpopulation appears to be genetically unique (Malenfant *et al.* 2016). The available population data are dated.

Sea ice habitat

There is no up-to-date sea-ice information. Throughout the 1990s, the preponderance of heavy multi-year ice through most of the central and western areas resulted in low densities of ringed seals (Kingsley *et al.* 1985) and, consequently, low densities of polar bears. However, if multi-year ice becomes more dynamic, habitat quality and productivity may improve over the short-term (Derocher *et al.* 2004; Markus *et al.* 2009; Sou and Flato 2009; Maslanik *et al.* 2011; Howell *et al.* 2008; Laidre *et al.* 2020).

Harvest management

The NW subpopulation is managed solely by Nunavut. The harvest quota for the NW subpopulation was set to 4 bears (3 males and 1 female) in 1996. The 5-year mean harvest (2015/16 – 2019/2020) of 1.2 bears/year is below a sustainable harvest level for that population size. The only Nunavut community that harvests from NW is Grise Fiord.

In September 2019, the new Nunavut Polar Bear Co-management Plan was approved by the Nunavut Wildlife Management Board, following five years of consultation and development with co-management partners in Nunavut. The Plan replaced the Memoranda of Understanding that had been in place. Concomitant with the approval of the new Plan, and in response to public and stakeholder feedback, Nunavut changed the allowable harvest sex-ratio. Beginning with the 2019/2020 harvest season, up to 50% of a community's quota can be harvested as females without entering an overharvest situation. This replaces the 2:1 male-to-female harvest sex-ratio and no changes to existing community TAHs were made. There is a potential that the biological risk of negative population outcomes due to harvest will increase because adult female polar bears are the most important contributors to population growth. The GN is monitoring how this implemented change in harvest sex-ratio will affect the sex-ratio of harvested bears. The average female proportion of harvest in 2019/2020 until 2021/2022 is 0%. Females have been underharvested relative to the annual recommended quota by approximately 80% on average during 2017/2018 – 2021/2022 (5 yr) time period. In February 2022, the Polar Bear Flexible Quota System was replaced by the Harvest Administration and Credit Calculation System. The up to 1:1 harvest sex-ratio is still in place.

References

- Derocher, A.E., Lunn, N.J., and Stirling, I. 2004. Polar bears in a warming climate. *Integrative and Comparative Biology* **44**:163–176.
- Howell, S.E.L., Tivy, A., Yackel, J.J., and McCourt, S. 2008. Multi-year sea-ice conditions in the Western Canadian Arctic Archipelago region of the Northwest Passage: 1968-2006. *Atmosphere-Ocean* **46**:229–242.
- Kingsley, M.C.S., Stirling, I., and Calvert, W. 1985. The distribution and abundance of seals in the Canadian High Arctic, 1980-82. *Canadian Journal of Fisheries and Aquatic Sciences* **42**:1189–1210.
- Laidre, K.L., Atkinson, S.N., Regehr, E. V, Stern, H.L., Born, E.W., Wiig, Ø., Lunn, N.J., Dyck, M., Heagerty, P. and Cohen, B.R. 2020. Transient benefits of climate change for a high-Arctic polar bear (*Ursus maritimus*) subpopulation. *Global Change Biology* **26**:6251–6265.
- Malenfant, R.M., Davis, C.S., Cullingham, C.I., and Coltman, D.W. 2016. Circumpolar genetic structure and recent gene flow of polar bears: a reanalysis. *PLoS ONE* **11**(3):e0148967, <https://doi.org/10.1371/journal.pone.0148967>.
- Markus, T., Stroeve, J.C., and Miller, J. 2009. Recent changes in Arctic sea ice melt onset, freezeup, and melt season length. *Journal of Geophysical Research* **114**:C12024, <https://doi.org/10.1029/2009JC005436>.
- Maslanik, J., Stroeve, J., Fowler, C., and Emery, W. 2011. Distribution and trends in Arctic sea ice through spring 2011. *Geophysical Research Letters* **38**:L13502, <https://doi.org/10.1029/2011GL047735>.
- Sou, T., and Flato, G. 2009. Sea ice in the Canadian Arctic Archipelago: modeling the past (1950-2004) and the future (2041-60). *Journal of Climate* **22**:2181–2198.
- Stirling, I. 1997. The importance of polynyas, ice edges and leads to marine mammals and birds. *Journal of Marine Systems* **10**:9–21.
- Stirling, I., Andriashek, D., and Calvert, W. 1993. Habitat preferences of polar bears in the western Canadian Arctic in late winter and spring. *Polar Record* **29**:13–24.
- Taylor, M.K., Laake, J., McLoughlin, P.D., Cluff, H.D., and Messier, F. 2008. Mark-recapture and stochastic population models for polar bears of the high Arctic. *Arctic* **61**:143–152.
- Taylor, M.K., Akeagok, S., Andriashek, D., Barbour, W., Born, E.W., Calvert, W., Cluff, H.D., Ferguson, S., Laake, J., Rosing-Asvid, A., Stirling, I., and Messier, F. 2001. Delineating Canadian and Greenland polar bear (*Ursus maritimus*) populations by cluster analysis of movements. *Canadian Journal of Zoology* **79**:690–709.

Southern Beaufort Sea (SB)

Boundary

The boundary for the Southern Beaufort (SB) subpopulation extends from 133°W at approximately Tuktoyaktuk, west to Icy Cape, Alaska. This boundary was formally accepted by management authorities for the Northern and Southern Beaufort Sea polar bear subpopulations in 2013. The previous boundary between the SB and Northern Beaufort Sea (NB) polar bear subpopulation existed at approximately 125°W longitude, near Pearce Point, NWT (Brower *et al.* 2002). The boundary change was proposed by researchers, resulting from radio telemetry studies that suggested this boundary did not reflect the space use patterns of bears in the eastern portion of the southern Beaufort Sea. Records indicated that approximately 90% of the bears harvested near Baillie Islands were actually NB bears (Amstrup *et al.* 2005). In consideration of the apparent misallocation of NB bears to the SB harvest, the WMAC (NWT) and IGC consulted with communities regarding the potential to change the SB/NB boundary. As a result, in 2013/14, the boundary was moved west to 133°W longitude, near the community of Tuktoyaktuk, NWT. The proportional representation of NB versus SB bears reduces to approximately 50:50 at this longitude, thus allowing harvest to be more accurately allocated between the subpopulations. A re-analysis of Regehr *et al.* 2001-2006 capture data was undertaken to estimate the SB and (NB) subpopulations under the new boundary (Griswold *et al.* 2017). The mean number of bears moved from the SB to NB is 311, which is being used until another subpopulation estimate is available (Griswold *et al.* 2017).

Overview of co-management partners and management objectives

Management of the SB subpopulation is jurisdictionally complex. In Canada, there is a co-management structure which involves the governments of Canada, the Northwest Territories, and Yukon as well as the Inuvialuit Game Council, the Wildlife Management Advisory Councils (NWT and North Slope) and the Inuvialuit Hunters and Trappers Committees. The SB subpopulation is shared with Alaska and cooperatively managed under the *Inuvialuit-Inupiat Polar Bear Management Agreement in the Southern Beaufort Sea*, originally signed in 1988 and subsequently revised. The harvest quota is recommended under the principles of this agreement by the designated Commissioners of the North Slope Borough and the Inuvialuit Game Council, and technical advisors. The primary management objectives in the *Inuvialuit-Inupiat Polar Bear Management Agreement in the Southern Beaufort Sea* are:

- To maintain a healthy viable population of polar bears in the southern Beaufort Sea in perpetuity, and
- To manage polar bears on a sustained yield basis in accordance with all the best information available whereby the acceptable annual harvest level does not exceed net annual recruitment to the population and accounts for all forms of removal from the population

Under the *Species At Risk (NWT) Act*, polar bears are listed as a species of Special Concern. The *Inuvialuit Settlement Region Polar Bear Joint Management Plan* was published in 2017 for the species; the goal of this plan is to ensure the long-term persistence of healthy polar bears in the ISR while maintaining traditional Inuvialuit use.

Indigenous Knowledge

To date, a number of Indigenous knowledge (IK) studies have been completed that cover the ISR polar bear populations; the largest in scope and most recent is the 2015 book *Inuvialuit and Nanuq: A polar bear traditional knowledge study*. It is important to note, however, there can be significant IK presented orally during events like project consultation meetings or public hearings that is often not adequately captured in a way to use as reference material. *Inuvialuit and Nanuq* describes Inuvialuit knowledge:

“The most important aspects of Indigenous knowledge concerning polar bears are the intergenerational knowledge (acquired from parents, grandparents and other elders) combined with direct experience. In general, this is what Inuvialuit mean by Traditional Knowledge (TK): personal knowledge acquired by travelling across ice, hunting seals and polar bears, running dog teams, reading wind directions, snow and cloud patterns, geographic features, currents and stars, and by intergenerational transmission.” – JS 2015, p. 9

Inuvialuit note that “ice conditions, the effects of climate change and polar bear behaviour are extremely complex.” (JS 2015, p. 197) Inuvialuit are reluctant to speculate about the future and long-term polar bear survival trends, given the high level of uncertainty in ecological conditions and how both bears and Inuvialuit will respond to these changes (JS 2015). Inuvialuit knowledge provides relative observations that can, in some cases, be used comparatively, to assess trends over time, or to draw a fuller picture of the SB polar bear subpopulation.

Abundance

During the verification workshop for *Inuvialuit and Nanuq* (JS 2015), consensus statements on changes to polar bear abundance over the lifetime of the TK holders were generated for each community:

- Tuktoyaktuk: “I would say they are the same. Overall throughout the years, they seem pretty stable. The bears are there, just a little bit later. It’s just the ice conditions that are changing” (JS 2015, p. 184)
- Aklavik: “I think I’m just too far away to see. But the talk around Aklavik is they are about the same numbers. I would agree with everyone” (JS 2015, p. 184)

Sources of Inuvialuit knowledge indicate that the polar bear population cycles over time, that bears tend to follow seals, and that an observed regional decline does not necessarily indicate a population decline. Inuvialuit knowledge holders have noted that seals are highly mobile and experience population cycles (JS 2015).

The 2021 *Species Status Report for Polar Bear (Ursus maritimus) in the Northwest Territories* (SARC, 2021) provides a summary of other sources of Inuvialuit and local knowledge of relative polar bear and seal abundance in the Inuvialuit Settlement Region (see that SARC 2021 for further detail). This historical perspective demonstrates the complexities of polar bear ecology and that the abundance and distribution of polar bears and their prey has always been variable.

Body condition

Inuvialuit knowledge holders in *Inuvialuit and Nanuq* (2015) agreed that polar bear body condition has remained generally stable over time, despite considerable variability within and between years. Inuvialuit also indicated that, since the 1980s, there have been less really big bears observed, and the big bears aren't as fat (JS 2015).

Distribution

Overall, Inuvialuit knowledge indicates that polar bear den locations have changed over time and bear sightings near Tuktoyaktuk have increased. Despite these observed changes in distribution, Inuvialuit assert that there is no evidence to suggest that these changes have had an impact on abundance (JS 2015). Inuvialuit knowledge holder interviews in Slavik *et al.* (2009) suggested that as sea ice melts in the southern Beaufort Sea, polar bear distribution will shift northward.

Climate Change

Inuvialuit see and experience climate change firsthand, noting changes to temperature, freeze-up, break-up, ice conditions, wind and storm patterns; the book *Inuvialuit and Nanuq* (2015) explores Inuvialuit observations of climate change since the 1980s. Despite observations of the climate change and a nuanced understanding of polar bear ecology, Inuvialuit have not yet seen changes to polar bear abundance or condition (JS 2015). Inuvialuit consensus is that:

“For the Inuvialuit, the future cannot be predicted; it could be good or bad as far as polar bears are concerned. However, the consensus among the workshop participants was that polar bears are highly intelligent animals that can adapt to climate change because they have been adapting to many things for thousands of years.” (JS, 2015 p. 196)

Scientific Knowledge

Abundance

Scientific study of the SB polar bear subpopulation extends back to the 1970s. It is worth noting that scientific polar bear research is very expensive and takes place in remote, dangerous areas where weather patterns are highly unpredictable. These factors can significantly affect study success and the frequency of population inventories.

The SB population declined substantially as harvest increased in the late 1950s/early 1960s due to sport hunting by non-aboriginal harvesters and increases in fur prices (Usher 1976, Amstrup *et al.* 1986, Amstrup 1995).

There have been multiple inventories conducted in the Southern Beaufort region, and all were based upon the former subpopulation boundaries. Results are summarized below:

Inventory period	Population Estimate	Confidence Intervals and Comments	Reference
1972-83	1,778	SD ± 803 CV=0.45	Amstrup <i>et al.</i> 1986
1992	Near 1,480		Amstrup 1995
1986-98	2,272 (2001)	Based on estimate of 1,250 females (C.V.=0.106); 55% females	Amstrup <i>et al.</i> 2001
2001-2006	1,526	95% CI=1211-1841; C.V.=0.106	Regehr <i>et al.</i> 2007
2006	1,215	Boundary change moves estimated 311 bears based on analysis in 2009.	Griswold <i>et al.</i> 2017

The current SB subpopulation estimate used for management is 1,215. This estimate is based on the Regehr *et al.* (2006) estimate (1,526) for the previous subpopulation area adjusted for new boundary at 133°W (Tuktoyaktuk) following a 2009 analysis by Griswold *et al.* (published in 2017), which indicated 311 bears would shift from the SB to the NB under the aforementioned boundary shift. This current SB subpopulation estimate cannot be compared with the historical estimates to assess trend. However, the Regehr *et al.* (2006) population estimate of 1,526 bears can be compared to the previous (Amstrup *et al.* 2001) population estimate of 2,272; the more recent estimate is lower, but the difference is not statistically significant.

Trends in Abundance and Body Condition

A recent population trend analysis by Bromaghin *et al.* (2015), suggests that a decline occurred in the SB polar bear abundance and body condition in the mid-2000s, coinciding with years of heavy sea ice conditions. The trend analysis suggests that abundance and body condition began to increase again towards the later 2000s. The study area and sampling regime on the Canadian side of the study area was inconsistent, which introduced bias into the analysis. It is difficult to assess the impact of this bias on the trend analysis. Further abundance and body condition trend analysis was completed for the Alaska side of the subpopulation (Atwood *et al.* 2020). This analysis found that abundance and body condition, in the Alaskan side of the SB, declined in the mid-2000s and then stabilized through to 2015.

Habitat/Climate Change

Scientific studies in the SB have noted sea ice declines (Durner *et al.* 2009), a reduction in body size and cub recruitment of SB bears in Alaska (Rode *et al.* 2010), and modeling that suggests declines in survival and breeding rates are related to increases in the ice-free period (Regehr *et al.* 2010). Multiple indicators of climate change impacts on sea ice have been noted for the SB polar bear subpopulation. From 1979 to 2014, researchers have observed: a declining number of ice-covered days, a declining rate of June to October sea ice concentration, and an increasing length of the summer season (Stern and Laidre 2016). The length of the summer season increased by 17.8 days from 1979 to 2014 for the SB (Stern and Laidre 2016).

Using fatty acid signature, the diet of polar bears in the Beaufort was dominated by ringed seal, but there was variation. Most variation in bear diet was explained by longitude, reflecting spatial variation in prey availability. Sea ice conditions (extent, thickness, and seasonal duration) declined throughout the study period, and date of sea ice break-up in the preceding spring was positively correlated with female body condition and consumption of beluga whale (Florko *et al.* 2020)

Future Abundance Work

The Commissioners to the *Inuvialuit-Inupiat Polar Bear Management Agreement in the Southern Beaufort Sea* have recommended non-invasive population survey methods for the SB subpopulation. In 2017 an aerial survey method was tested, but failed to produce a robust population estimate. A genetic mark-recapture survey was designed instead, encompassing the SB and NB subpopulations. The first year of field work was 2019, with three to four years of fieldwork planned in total. The 2020 field season was postponed due to the Covid-19 pandemic, but resumed in 2021. Discussions are ongoing about the possible inclusion of additional marks for this survey using fecal samples and eDNA from polar bear tracks. In 2020, the Joint Secretariat hired contractors to develop a potential framework for incorporating IK into an Integrated Population Model (IPM) for SB and NB polar bears, with a focus on incorporating IK as a source of information..

Harvest Management

Within the ISR, harvest is carefully managed. All human-caused mortality including hunting, defense of life and property kills, industry-related mortalities and illegal kills are tracked and counted under a quota. There is mandatory reporting and submission of proof of sex and age that is enforceable under the Wildlife Act. In Alaska, the Southern Beaufort harvest has been under an effective voluntary quota since 1988 and is currently monitored by the North Slope Borough and USFWS through a marking, tagging, and reporting program (USFWS 2010). A key aspect that ensures human caused mortality remains below TAH is a highly adaptive management system whereby information related to population abundance and trend is evaluated annually by the WMACs (NWT and NS) and the IGC and changes are recommended to HTC's or the Minister when required.

Hunting in the SB was historically largely conducted by non-Indigenous harvesters (Usher 1976). Quotas were first applied in Canada for the 1967-68 hunting season. In the absence of data, quotas for each settlement were established by averaging the harvest of the previous 3 years and then reducing that number by a modest amount (Brower *et al.* 2002). The first quota increases based on scientific information were made in 1978-79 after completion of the first population study of polar bears in the Western Arctic (Stirling *et al.* 1975). Quotas were based on the understanding that the total harvest of independent females would not exceed the modelled sustainable maximum of 1.5% of the population (Taylor *et al.* 1987) and that a 2:1 ratio of males to females would be maintained in the total quota harvested (Stirling 2002).

Currently, Inupiat and Inuvialuit have exclusive rights to harvest polar bears from the SB. The *Inuvialuit-Inupiat Polar Bear Management Agreement in the Southern Beaufort Sea* (signed 1988, revised 2011) sets out management principles including agreed upon harvest limits. These harvest quotas are mandatory in Canada. Since the signing of the agreement the quota has ranged from a maximum of 80 to the current quota of 56 with the new boundary. In recent years the entire quota is rarely taken (see Indigenous Knowledge, Harvest). Harvest of the SB subpopulation has been below the quota for several years. Changing sea ice, distance needed to travel, challenging ice conditions and cost of travel to access bears are all cited as reasons (Larry Carpenter pers. comm. 2020). Changing sea ice conditions has made it difficult for Inuvialuit to rely on established IK for planning harvest activities (JS, 2015).

Protected areas

All denning habitat along the Yukon coast is protected by Ivvavik National Park, Herschel Island Qikiqtaruk Territorial Park and the land withdrawal on the Eastern Yukon North Slope. In Alaska, a large proportion of the coast is protected by 1002 lands in Arctic National Wildlife Refuge, however, recent changes in US law has potentially opened this area to oil and gas. The future of oil and gas development in the 1002 lands, which includes critical polar bear denning habitat for Southern Beaufort polar bears, is unclear as of February 2022. In Canada bears in dens are protected by Hunters and Trappers Committee by-laws and regulations. The Inuvialuit-Inupiat agreement also protects bears in dens.

References

- Amstrup, S. C., I. Stirling, and J. W. Lentfer. 1986. Past and present status of polar bears in Alaska. *Wildlife Society Bulletin* 14:241-254.
- Amstrup, S. C. 1995. Movements, distribution, and population dynamics of polar bears in the Beaufort Sea. Ph.D. Dissertation. University of Alaska-Fairbanks, Fairbanks, AK. 299 pp.
- Amstrup, S.C., T.L. McDonald and I. Stirling. 2001. Polar bears in the Beaufort Sea: a 30- year mark-recapture case history. *J. Agri., Bio., and Environ. Stats.* 6(2):221-234.
- Amstrup, S.C., G.M. Durner, I. Stirling and T.L. McDonald. 2005. Allocating harvest among polar bear stocks in the Beaufort Sea. *Arctic* 58(1):247-259.
- Atwood, T. C., Bromaghin, J. F., Patil, V. P., Durner, G. M., Douglas, D. C. and Simac, K. S. 2020. Analyses on subpopulation abundance and annual number of maternal dens for the U.S. Fish and Wildlife Service on polar bears (*Ursus maritimus*) in the Southern Beaufort Sea, Alaska: U.S. Geological Survey Open-File Report 2020-1087, <https://doi.org/10.3133/ofr20201-87>.

- Bromaghin, J. F., McDonald, T. L., Stirling, I., Derocher, A. E., Richardson, E. S., Regehr, E. V., Douglas, D. C., Durner, G. M., Atwood, T. and Amstrup, S. C. 2015. Polar bear population dynamics in the southern Beaufort Sea during a period of sea ice decline. *Ecological Applications*, 25: 634-651. doi:10.1890/14-1129.1
- Brower, C.D., A. Carpenter, M.L. Branigan, W. Calvert, T. Evans, A.S. Fischbach, J.A. Nagy, S. Schliebe, and I. Stirling. 2002. The Polar Bear Management Agreement for the Southern Beaufort Sea: An evaluation of the first ten years of a unique conservation agreement. *Arctic* 55:362–372.
- Canadian Wildlife Service (CWS). 2010. Northwest Territories – Inuvialuit Consultation Report on Polar Bear: Report on the Consultation in the Northwest Territories Inuvialuit Settlement Region in February 2009 on the Proposed Listing of Polar Bear as a Species of Special Concern Under the Federal Species at Risk Act. Canadian Wildlife Service, Yellowknife, NT. 121 pp.
- Carpenter, L. 2020. Personal communication at the Polar Bear Technical Committee meeting, February 3-6, 2020.
- Durner, G. M., D. C. Douglas, R. M. Nielson, S. C. Amstrup, T. L. McDonald, I. Stirling, M. Mauritzen, E. W. Born, Ø. Wiig, E. DeWeaver, M. C. Serreze, S. E. Belikov, M. M. Holland, J. Aars, D.A. Bailey and A.E.Derocher. 2009 Predicting 21st-century polar bear habitat distribution from climate models. *Eco. Mons.* 79(1):25-58. <https://doi.org/10.1890/07-2089.1>
- Florko, K.R. G.W. Thiemann and J.F. Bromaghin. 2020. Drivers and consequences of apex predator diet composition in the Canadian Beaufort Sea. *Oecologia*. 194:51-63.
- Griswold J., T. McDonald, M. Branigan, E. V. Regehr, and S. C. Amstrup. 2017. Southern and northern Beaufort Sea polar bear population estimates under a proposed boundary shift. Manuscript Report 265.
- Joint Secretariat. 2015. Inuvialuit and Nanuq: A Polar Bear Traditional Knowledge Study. Joint Secretariat, Inuvialuit Settlement Region. xx + 304 pp. Available online: http://www.wmacns.ca/pdfs/394_polar-bear-tk-report-low-res.pdf.
- Joint Secretariat. 2017. Inuvialuit Settlement Region Polar Bear Joint Management Plan. Joint Secretariat, Inuvialuit Settlement Region. vii + 66 pp. https://www.nwtspciestrisk.ca/sites/default/files/isr_polar_bear_joint_management_plan_2017_final.pdf
- Regehr, E. V., S. C. Amstrup, and I. Stirling. 2007. Polar Bear Population Status in the Southern Beaufort Sea. U.S. Geological Survey Open-File Report 2006-1337. 20 pp.
- Regehr, E. V., C. M. Hunter, H. Caswell, S. C. Amstrup, and I. Stirling. 2010. Survival and breeding of polar bears in the southern Beaufort Sea in relation to sea ice. *Journal of Animal Ecology* 79:117-127.
- Rode, K. D., S. C. Amstrup, and E. V. Regehr. 2010. Reduced body size and cub recruitment in polar bears associated with sea ice decline. *Ecological Applications* 20:768-782.
- Species at Risk Committee. 2012. Species Status Report for Polar Bear (*Ursus maritimus*) in the Northwest Territories. Species at Risk Committee, Yellowknife, NT.
- Slavik, D., Wildlife Management Advisory Councils (Northwest Territories and North Slope), and Inuvialuit Game Council. 2009. Inuvialuit Knowledge of Nanuq: Community and Traditional Knowledge of Polar Bears in the Inuvialuit Settlement Region. Inuvialuit Joint Secretariat, Inuvik, NWT. 67 pp. Report and Unpublished Transcripts.
- Stirling, I., D. Andriashek, P. Latour, and W. Calvert. 1975. The distribution and abundance of polar bears in the eastern Beaufort Sea. Final Report to the Beaufort Sea Project. Fisheries and Marine Service, Department of Environment, Victoria, B.C. 59 pp.
- Stirling, I. 2002. Polar bears and seals in the eastern Beaufort Sea and Amundsen Gulf: A synthesis of population trends and ecological relationships over three decades. *Arctic* 55 (Supp. 1):59–76.
- Taylor, M.K., DeMaster, D.P., Bunnell, F.L. and Schweinsburg, R.E. 1987. Modeling the sustainable harvest of polar bears. *J. Wildl. Manage.* **51**:811-820.
- Usher, P. 1976. Inuit Land Use in the Western Canadian Arctic. Pp. 21-31 in: Inuit Land Use and Occupancy Report, Vol. 1. M. M. R. Freeman, ed. Department of Indian Affairs and Northern Development, Ottawa, ON.

Southern Hudson Bay (SH)

Status and Delineation

Boundaries of the Southern Hudson Bay polar bear subpopulation are based on observed movements of marked and collared bears (Jonkel *et al.* 1976; Kolenosky and Prevett 1983; Kolenosky *et al.* 1992; Obbard and Middel 2012; Middel 2013). The range of the SH subpopulation includes much of eastern and southern Hudson Bay and James Bay and large expanses of the coastline of Ontario and Québec as well as areas up to 120 km inland (Kolenosky and Prevett 1983; Obbard and Walton 2004; Obbard and Middel 2012). Inuit Knowledge has indicated that, in Nunavik, there were very few bears from the 1940s to the 1960s, with somewhat of an increase in the population from the 1960s to the 1980s, and a marked increase since the 1980s (NMRWB 2018).

The first population estimate for SH came from a three-year (1984–1986) mark-recapture study, conducted mainly along the coastline and inland areas of Ontario (Kolenosky *et al.* 1992). The initial estimate obtained from that study (763 ± 323 bears) was later corrected to 641 bears (95% CI: 401 – 881) after a re-analysis of the original capture data (Obbard *et al.* 2007) but covered only the area along and inland from the Hudson Bay coastline in Ontario and did not cover other jurisdictions, the area around James Bay, south of Hook Point or Akimiski Island. A subsequent 3-year capture-recapture study (2003–2005), covering this same area, but likely more comprehensively covering the inland areas of Ontario, produced an estimate of 681 bears (95% CI: 401–961; Obbard *et al.* 2007). An analysis of bears captured on Akimiski Island in James Bay during 1997 and 1998 resulted in the addition of 70–110 bears (Obbard *et al.* 2007) and the total SH subpopulation was therefore considered by the PBTC to be between 900–1000 bears for management purposes. Results from the two capture-recapture studies suggested that the abundance was unchanged between 1984–1986 and 2003–2005, though survival rates and body condition in all age and sex categories declined (Obbard 2008; Obbard *et al.* 2016). Inuit Knowledge from Nunavik, further north in the subpopulation's range, indicated a very large increase in observations of bears at this time, and no apparent declines in health (NMRWB 2018).

An aerial survey was conducted during the fall ice-free season over mainland Ontario (same geographic area as for the capture–recapture studies plus all areas along and inland from the coast of James Bay) and Akimiski Island in 2011 and over the remaining islands in James Bay, the coastal areas of Québec from Long Island to the SH–FB subpopulation management boundary, and the off-shore islands in eastern Hudson Bay in 2012. Results of this combined mark-recapture distance-sampling (MRDS) and double-observer mark-resight survey provided an estimate of 860 bears (95% CI: 580–1,274) on the mainland of Ontario, neighboring islands, and Akimiski Island portions of the SH management unit during the 2011 ice-free season, plus an additional 83 bears (SE = 4.5) in the 2012 study area. Thus, combining the aerial survey results from 2011 and 2012 yielded an overall estimate of 943 bears (SE: 174, 95% CI: 658–1350) for SH (Obbard *et al.* 2015). Overall, despite the difference in methodologies, assumptions, and biases between capture–recapture studies and aerial surveys, these lines of evidence suggest it is likely that the subpopulation had not substantially changed in abundance between the mid-1980s and 2012. Nevertheless, the duration of sea ice within the bounds of SH declined over this period (Hochheim and Barber 2014; Stern and Laidre 2016; NMRWB 2018) and scientific research also indicated a decline in body condition of bears during that same period (Obbard *et al.* 2016). Nunavik Inuit Knowledge indicated there may have been a population increase during this time and unchanging good health (NMRWB 2018).

An aerial survey, covering the same areas as the 2011/12 survey, was repeated in September 2016 to re-assess the abundance in SH. All areas in Ontario, Nunavut and Québec were sampled within a 3-week period to ensure complete coverage within the same season and year. The abundance estimate obtained from that survey (780 bears, 95% CI: 590–1029) suggested that the subpopulation had declined by approximately 17% between 2011/12 and 2016. The proportion of yearlings in the observed portion of the subpopulation also declined from 12% in 2011 to 5% in 2016, whereas the proportion of cubs remained similar (16% in 2012 vs. 19% in 2016), suggesting a low survival of cubs to yearlings (Obbard *et al.* 2018). Inuit knowledge from Nunavik indicated that the number of bears being sighted was among the highest it had been in a lifetime at the time of data collection in late 2014 and early 2015, although there were (sometimes very notable) fluctuations from year to year (NMRWB 2018).

To assess if the apparently low survival rate of cubs observed during the 2016 survey was an unusual event or represented an ongoing trend for SH, a partial survey of the Ontario coastline was conducted in September 2018. The results of this survey indicated a slightly lower abundance in the coastal area in 2018 (249 bears, 95% CI: 230 – 270) than in 2016 (269 bears, 95% CI: 244 – 297) and significantly lower abundance than in 2011 (422 bears, 95% CI: 381 – 467). The proportion of yearlings in the coastal area for the three surveys was variable (2011: 12%, 2016: 3%, 2018: 7%) as was the number of cubs (2011: 15%, 2016: 17%, 2018: 10%), but the proportion of adults in the coastal area increased in each survey (2011: 60%, 2016: 71%, 2018: 74%). The results of the 2018 survey should be used tentatively, as they are not a complete sample of the subpopulation.

A third comprehensive aerial survey was flown in Summer 2021 to update the abundance estimate and determine trend of the SH subpopulation (Northrup *et al.* 2022). Two separate estimates were derived: N=1003 (95% CI: 773–1302) that assumed perfect detection on the transect line as Obbard *et al.* (2018) to allow for direct comparison to the 2016 survey, and N=1119 (95% CI: 860–1454) that took advantage of a novel approach to estimate the probability of detection on the transect line while accounting for the blind spot affecting rear observers (Wiig *et al.* 2022). The former estimate was most comparable to the 2016 estimate, but the latter was a more robust estimate of the true subpopulation size in 2021. Both estimates indicate a greater number of bears within the boundaries of the subpopulation than in 2016. The first estimate (N=1003), which was directly comparable to Obbard *et al.* (2018), suggested a 29% increase. On the 2021 survey, 148 family groups were observed, including those seen while off transect or transiting; 75 were females with cubs and 73 with yearlings. The average cub and yearling litter size was 1.57 and 1.47, respectively. Cubs comprised 16%, 19% and 14% of bears in SH in 2011, 2016 and 2021 and yearlings comprised 12%, 5% and 13% of observed bears (Obbard *et al.* 2015; Obbard *et al.* 2018; Northrup *et al.* 2022). This increase in total abundance of bears within the boundaries of SH was likely driven by a combination of improved vital rates and interannual variation in the on-land distribution of bears in the adjacent Western Hudson Bay polar bear subpopulation. Between 2016 and 2021, reported harvest in SH was 37.8 bears per year, which was substantially lower than between 2010 and 2015 (58.8 bears per year). Further, the 3 longest ice seasons between 2010 and 2021

occurred between 2016 and 2021. These facts, suggest some improvement in demographic rates was likely. However, genetic mark-recapture surveys taking place in WH and SH between 2018 and 2021 indicated that 25% of the bears sampled in SH in 2021 (overlapping temporally with the survey), that were previously marked either via physical capture or biopsy darting were previously only ever sampled in WH (McGeachy *et al.* 2023). Mcgeachy et al. (2023) further assessed annual movement of bears between 2021 and 2022 using animals biopsy darted in both years. They found that 93% of bears sampled in WH in 2021 and resampled in 2022 were still in WH, with the other 7% having moved to SH. Contrastingly, they found that only 64% of bears sampled in SH in 2021 and resampled in 2022 were still in SH, with the remaining 36% being resampled in WH. Collectively, these results suggest substantial movement of bears into SH from WH in 2021, with subsequent movements back to WH.

Harvest Management

The SH subpopulation range overlaps Nunavut, Québec and Ontario. A Total Allowable Harvest (TAH) of 25 bears is currently in place for Sanikiluaq which is the only Nunavut community harvesting within SH. A Total Allowable Take (TAT) of 23 bears is in place in the Nunavik Marine Region, for Inuit and Cree from Québec, with a minimum of 1 bear allocated to the Cree. Part of the range occupied by the SH subpopulation (inland Québec, Ontario and part of the Eeyou Marine Region) does not have a harvest limitation.

An interjurisdictional process is underway for the entire SH subpopulation area to reassess the TAT/TAH for SH with regards to new scientific and IK information. As part of this process, a harvest risk assessment for SH was completed (Regehr *et al.* 2021), which suggested that this subpopulation has historically had a strong capacity for growth and was likely able to sustain harvest at relatively high rates (compared to other subpopulations). This assessment examined a range of possible scenarios for how polar bears in SH might respond to ongoing climate change and how this might influence the sustainability of harvest. Under what the authors considered to be a moderate level of influence of climate change on population growth, similar to what this subpopulation experienced over the last 15 years, the risk assessment indicated that the subpopulation could sustain a 3-4.5% harvest rate at a harvest sex ratio of 2 males to 1 female or a 2-3% harvest rate at a harvest sex ratio of 1 male to 1 female while still maintaining the population abundance above the maximum net productivity level.

While the current total potential harvest (>48 bears) is above what would be considered sustainable under these conditions, reported harvest over the last decade falls within the bounds suggested by the risk assessment study. However, considering that harvest reporting is incomplete in Ontario and Québec, current estimates of harvest are likely biased low. The Québec government, in collaboration with the Nunavik and Eeyou Marine Regions Wildlife Boards, has recently adopted a first polar bear management plan which applies to both the inland and marine portions areas where polar bears are harvested by communities in Nunavik and Eeyou Istchee. This plan provides a framework for the management of polar bear subpopulations occurring within Québec, including SH, and aims at achieving complete reporting of human-caused mortalities of polar bears. In 2022, Sanikiluaq applied for 6 credits (1 male and 5 females) under the Nunavut harvest management system from a total of 11 available credits (5.03 female and 5.97 male) that had accrued under Nunavut's flexible quota system. The resulting harvest from Sanikiluaq was 31 bears, and the total reported harvest from all jurisdictions was 37 bears, which represents a 4.7% and 3.3% harvest rate of the 2016 and 2021 population estimates, respectively. To date, credits existing under the Nunavut harvest management system have never been reset for this subpopulation.

In recent years, discussions with Indigenous communities around James Bay in Ontario and Québec suggest that bears are being encountered and coming into conflict with people on the land and in communities more regularly. Reported defense of life and property kills during 2018-2019 and 2019-2020 in Ontario were also at their highest level since the 2007-2008 season but were lower in 2020-2021 and 2021-2022. Land users in the Eeyou Marine Region are also reporting an increase in conflicts on the islands in southern James Bay, and to a lesser extent, along the East Coast of the Bay.

References

- Hochheim, K.P., and Barber, D.G. 2014. An update on the ice climatology of the Hudson Bay system. *Arctic, Antarctic, and Alpine Research* **46**:66–83.
- Jonkel, C., Smith, P., Stirling, I., and Kolenosky, G.B. 1976. The present status of the polar bear in the James Bay and Belcher Islands area. Canadian Wildlife Service Occasional Paper 26, 42 pp.
- Kolenosky, G.B., and Prevett, J.P. 1983. Productivity and maternity denning of polar bears in Ontario. *International Conference for Bear Research and Management* **5**:238–245.
- Kolenosky, G.B., Abraham, K.F., and Greenwood, C.J. 1992. Polar bears of southern Hudson Bay. Polar bear project, 1984-88. Final Report. Ontario Ministry of Natural Resources, Ontario, Canada, 107 pp.
- McGeachy, D., Lunn, N.J., Northrup, J.M., Trim, V., Davis, C., and Derocher, A.E. 2023. Distributional shifts of polar bears (*Ursus maritimus*) in Hudson Bay in relation to sea ice dynamics, 2017-2022 Final report. Environment and Climate Change Canada, 32 pp.
- Middel, K.R. 2013. Movement parameters and space use for the Southern Hudson Bay polar bear subpopulation in the face of a changing climate. M.Sc. thesis, Trent University, Peterborough, Ontario, Canada.
- NMRWB [Nunavik Marine Region Wildlife Board]. 2018. Nunavik Inuit Knowledge and observations of polar bears: Polar bears of the Southern Hudson Bay sub-population. Project conducted and report prepared for the NMRWB by Basterfield, M., Breton–Honeyman, K., Furgal, C., Rae, J., and O' Connor, M. Available at: <https://nmrwb.ca/wp-content/uploads/2017/05/NMRWB-Nunavik-Inuit-knowledge-and-Observations-of-polar-bear-s-SHBsubpopulation.pdf>.
- Northrup, J.M., Howe, E., Lunn, N.J., Middel, K., Obbard, M.E., Ross, T.R., Szor, G., Walton, L., and Ware, J. 2022. 2021 Southern Hudson Bay polar bear subpopulation aerial survey: report to the 2023 meeting of the Canadian Federal-Provincial-Territorial

- Polar Bear Technical Committee, Québec City, Québec, 31 January–3 February. Unpublished Typescript, Ontario Ministry of Natural Resources, Peterborough, 52 pp.
- Obbard, M.E. 2008. Southern Hudson Bay polar bear project 2003–2005: Final report. Unpublished report, Wildlife Research and Development Section, Ontario Ministry of Natural Resources, Peterborough, Ontario, Canada, 64 pp.
- Obbard, M.E., Cattet, M.R.L., Howe, E.J., Middel, K.R., Newton, E.J., Kolenosky, G.B., Abraham, K.F., and Greenwood, C.J. 2016. Trends in body condition in polar bears (*Ursus maritimus*) from the Southern Hudson Bay subpopulation in relation to changes in sea ice. *Arctic Science* **2**:15–32.
- Obbard, M.E., McDonald, T.L., Howe, E.J., Regehr, E.V., and Richardson, E.S. 2007. Polar bear population status in southern Hudson Bay, Canada. U.S. Geological Survey Administrative Report, U.S. Department of the Interior, Reston, Virginia, USA, 34 pp.
- Obbard, M.E., and Middel, K.R. 2012. Bounding the Southern Hudson Bay polar bear subpopulation. *Ursus* **23**:134–144.
- Obbard, M.E., Stapleton, S., Middel, K.R., Thibault, I., Brodeur, V., and Jutras, C., and Dyck, M. 2015. Estimating the abundance of the Southern Hudson Bay polar bear subpopulation with aerial surveys. *Polar Biology* **38**:1713–1725.
- Obbard, M.E., Stapleton, S., Szor, G., Middel, K.R., Jutras, C., and Dyck, M. 2018. Re-assessing abundance of Southern Hudson Bay polar bears by aerial survey: effects of climate change at the southern edge of the range. *Arctic Science* **4**:634–655. [dx.doi.org/10.1139/as-2018-0004](https://doi.org/10.1139/as-2018-0004)
- Obbard, M.E., and Walton, L.R. 2004. The importance of Polar Bear Provincial Park to the Southern Hudson Bay polar bear population in the context of future climate change. Pages 105–116 in Rehbein, C.K., Nelson, J.G., Beechey, T.J., and Payne, R.J. (eds.) *Parks and Protected Areas Research in Ontario, 2004: Planning Northern Parks and Protected areas Areas. Proceedings of the Parks Research Forum of Ontario (PRFO) Annual General Meeting, May 4–6, 2004*. Parks Research Forum of Ontario, Waterloo, Ontario, Canada.
- Regehr, E.V., Dyck, M., Iverson, S., Lee, D.S., Lunn, N.J., Northrup, J.M., Richer, M.-C., Szor, G., and Runge, M. 2021. Incorporating climate change in a harvest risk assessment for polar bears *Ursus maritimus* in Southern Hudson Bay. *Biological Conservation* **258**, <https://doi.org/10.1016/j.biocon.2021.109128>.
- Stern, H.L., and Laidre, K.L. 2016. Sea-ice indicators of polar bear habitat. *The Cryosphere* **10**:2027–2041.
- Wiig, Ø., Atkinson, S.N., Born, E.W., Stapleton, S., Arnold, T., Dyck, M., Laidre, K.L., Lunn, N.J., and Regehr, E.V. 2022. An on-ice aerial survey of the Kane Basin polar bear (*Ursus maritimus*) subpopulation. *Polar Biology* **45**:89–100.

Viscount Melville Sound (VM)

Boundary

The Viscount Melville Sound subpopulation (VM) extends from northern Victoria Island through the Viscount Melville Sound to north of Melville Island, and from eastern M'Clure Strait, north to eastern Prince Patrick Island. The majority of the subpopulation area is within the Inuvialuit Settlement Region (ISR), with the eastern portion in Nunavut. A five-year study of movements and subpopulation size, using telemetry and mark-recapture, was completed for polar bears inhabiting VM in 1992 (Messier *et al.* 1992, 1994, Taylor *et al.* 2002). Population boundaries were based on observed movements of female polar bears with satellite radio-collars and movements of bears tagged in and out of the study area (Bethke *et al.* 1996, Taylor *et al.* 2001).

Overview of management partners and management objectives

The management partners and collaborating agencies for the VM subpopulation on the ISR side are the Government of the Northwest Territories, the WMAC (NWT), the Inuvialuit Game Council and Environment and Climate Change Canada. In Nunavut, management partners include Government of Nunavut, Ekaluktutiak Hunters and Trappers Organization, Kitikmeot Regional Wildlife Board and the Nunavut Wildlife Management Board. Management objectives and guiding principles for the VM and NB shared subpopulations are outlined in the *Polar Bear Management Agreement for the North[ern] Beaufort Sea and Viscount Melville Sound Polar Bear Populations between Inuit of the Kitikmeot West Region in Nunavut and the Inuvialuit* (2006). The key objectives of this agreement are:

- To maintain the Northern Beaufort Sea and Viscount-Melville Sound polar bear populations at healthy viable levels in perpetuity, and
- To manage polar bears on a sustained yield basis in accordance with all the best information available

Under the *Species at Risk (NWT) Act*, polar bears are listed as a species of Special Concern. The *Inuvialuit Settlement Region Polar Bear Joint Management Plan* was published in 2017 for the species; the goal of this plan is to ensure the long-term persistence of healthy polar bears in the ISR while maintaining traditional Inuvialuit use.

Indigenous Knowledge

To date, a number of Indigenous knowledge (IK) studies have been completed that cover the ISR polar bear populations; the largest in scope and most recent is the 2015 book *Inuvialuit and Nanuq: A polar bear traditional knowledge study*. It is important to note, however, there can be significant IK presented orally during events like project consultation meetings or public hearings that is often not adequately captured in a way to use as reference material. *Inuvialuit and Nanuq* describes Inuvialuit knowledge:

“The most important aspects of Indigenous knowledge concerning polar bears are the intergenerational knowledge (acquired from parents, grandparents and other elders) combined with direct experience. In general, this is what Inuvialuit mean by Traditional Knowledge (TK): personal knowledge acquired by travelling across ice, hunting seals and polar bears, running dog teams, reading wind directions, snow and cloud patterns, geographic features, currents and stars, and by intergenerational transmission.” – JS 2015, p. 9

Inuvialuit note that “ice conditions, the effects of climate change and polar bear behaviour are extremely complex.” (JS 2015, p. 197) Inuvialuit are reluctant to speculate about the future and long-term polar bear survival trends, given the high level of uncertainty in ecological conditions and how both bears and Inuvialuit will respond to these changes (JS 2015). Inuvialuit knowledge provides relative observations that can, in some cases, be used comparatively, to assess trends over time, or to draw a fuller picture of the VM polar bear subpopulation.

Abundance

During the verification workshop for *Inuvialuit and Nanuq* (JS 2015), consensus statements on changes to polar bear abundance over the lifetime of the TK holders were generated for each community:

- Ulukhaktok — “maybe a little change, but overall about the same. Polar bear movements are always different every year. To me it’s the same, but a little bit change since when I was younger” (JS 2015, p. 184)

In 2009, polar bear abundance was considered to be high around Melville Island (Slavik *et al.* 2009) and in 2009-2013, Inuit knowledge indicated that this subpopulation was increasing, based on information provided at the Canadian Wildlife Service Nunavut consultation meetings in 2009 (CWS unpublished) and community consultations in Cambridge Bay and Ulukhaktok during 2012 and 2013 (GNWT unpublished meeting notes).

Body Condition

Inuvialuit knowledge holders in *Inuvialuit and Nanuq* (JS 2015) agreed that polar bear body condition has remained generally stable over time, despite considerable variability within and between years. Inuvialuit also indicated that, since the 1980s, there have been fewer really big bears observed, and the big bears aren’t as fat (JS 2015).

Distribution

Overall, Inuvialuit knowledge indicates that polar bear den locations have changed over time and timing of females with cubs entering and leaving dens has changed. Despite these observed changes in distribution, Inuvialuit assert that there is no evidence to suggest that these changes have had an impact on abundance (JS 2015).

Climate Change

Inuvialuit see and experience climate change firsthand and have a nuanced understanding of polar bear ecology. Inuvialuit have not yet seen climate-related changes to polar bear abundance or condition (JS 2015). Inuvialuit consensus is that:

“For the Inuvialuit, the future cannot be predicted; it could be good or bad as far as polar bears are concerned. However, the consensus among the workshop participants was that polar bears are highly intelligent animals that can adapt to climate change because they have been adapting to many things for thousands of years.” (JS 2015, p. 196).

Scientific Knowledge

Abundance

Scientific study of the VM subpopulation extends back to the 1970s. It is worth noting that polar bear scientific research is very expensive and takes place in remote, dangerous areas where weather patterns are highly unpredictable. These factors can significantly affect study success and the frequency of population inventories.

The first subpopulation inventory for VM was conducted between 1989 and 1992 and yielded an estimate of 161 bears (SE = 34) (Taylor *et al.* 2002). There had been previous work (1974-1976) in the southern portion of the subpopulation area (Hadley Bay and Wynniatt Bay) as part of a broader study; however, no specific VM estimate was produced (Schweinsburg *et al.* 1981). Following fieldwork from 1989-1992, there was a concern that relatively high harvest rates and strong selection for males that occurred prior to the inventory had reduced the number of adult males in the population, impacting productivity. As a result, beginning in 1994, there was a five-year moratorium on harvest of VM bears. A subsequent simulation analysis using RISKMAN suggested that in 1999 (following the five-year moratorium) there was an estimated population of 215 (SE = 57.4) (Taylor *et al.* 2002). A subpopulation estimate for the VM has been submitted for publication (fieldwork conducted 2012-2014). Preliminary peer reviewed results, using a multi-state model that includes telemetry and harvest return data, estimates the 2014 population at 235 (95% CRI = 148-569 bears). This subpopulation estimate contrasts with the ‘superpopulation’ estimate (N=340, 95% CRI = 214-815) in the submitted manuscript that more accurately reflects the number of polar bears spending the majority of their time within the subpopulation boundary, rather than the total number of bears that spend even a short period of time within the area. The recent analysis suggests that the VM subpopulation has increased from an earlier period of decline and may have reached stability. Mean estimates of survival increased from 1989-1992 to 2012-2014 for all sex and age classes, except for adult females which declined. It is not known why adult female survival declined.

Habitat/Climate Change

Throughout the 1990s, the preponderance of heavy multi-year ice through most of the central and western areas resulted in low densities of ringed seals (Kingsley *et al.* 1985) and, consequently, low densities of polar bears. Multiple indicators of climate change impacts on sea ice have been noted for the VM polar bear subpopulation. From 1979 to 2014, researchers have observed: a declining number of ice-covered days, a declining rate of June to October sea ice concentration, and an increasing length of the summer season (Stern and Laidre 2016). The length of the summer season increased by 11.8 days from 1979 to 2014 for the VM (Stern and Laidre 2016). With the occurrence of multi-year ice shifting to more annual ice as compared to previous decades (e.g. Howell *et al.* 2015), habitat quality could be improved over the short-term. Increased open water in the summer would allow whales and other species to use the VM area, and this is reflected in the fatty acid diet signature in the VM. While ring seals remain the majority prey species, recent work showed VM bears with a higher proportion of beluga whale in their diet (37%) than any other subpopulation studied (Florko *et al.* 2021). This work also indicated lighter sea ice conditions were associated lower consumption of ringed seal in VM and body condition of VM bears were poorer than NB bears.

Harvest Management

Within the ISR, harvest is carefully managed. All human-caused mortality including hunting, defense of life and property kills, industry-related mortalities and illegal kills are tracked and counted under a quota. There is mandatory reporting and submission of proof of sex and age that is enforceable under the Wildlife Act. A key aspect that ensures human caused mortality remains below TAH is a highly adaptive management system whereby information related to population abundance and trend is evaluated annually by the WMACs (NWT and NS) and the IGC and changes are recommended to HTC and the Minister when required.

Within Canada, quotas were first established in NWT by the 33rd Session of the Territorial Council at Resolute Bay. The quotas were to become effective on July 1 for the 1967-68 hunting season. In the absence of data, quotas for each settlement were established by averaging the harvest of the previous 3 years and then reducing that number by a modest amount (Brower *et al.* 2002).

In 1973-74, the GNWT created a quota of 12 bears for Melville Island and 4 for Hadley Bay on northeast Victoria Island. Arguments (excerpts from PBTC minutes) supporting the establishment of this quota were: a) that it would be an added incentive for people to travel further from the settlements, particularly in years of fox abundance, b) a limited kill would allow accumulation of some information about the bear population in the area, which was currently lacking and, c) the kill would not cause irreparable damage and might give incentive for biological research in the area. At the time the PBTC suggested that the harvest should be monitored, along with full collection of specimens, and subject to review in due course when research had been conducted in the area.

Initially, the Hadley Bay quota was to be taken by hunters from Cambridge Bay. In 1980-81, the Hadley Bay quota was increased to 8. After the signing of the *Inuvialuit Final Agreement* (1984), Ulukhaktok began taking up to 8 of their community quota in Wynniatt Bay.

Although the Melville quota was hunted most often by Sachs Harbour and Ulukhaktok, it was also allocated to hunters from Resolute and other areas in the eastern Arctic. In 1984, the Melville quota was permanently assigned to be shared between Sachs Harbour and Ulukhaktok.

Beginning in the 1991-92 season, the quotas for Hadley Bay and Melville Island (8 and 12 respectively) were eliminated. Instead, Sachs Harbour, Ulukhaktok, and Cambridge Bay received an additional six tags each. The six bear allocations to Ulukhaktok and

Cambridge Bay were still allowed to be taken from Viscount-Melville Sound for 1991/92 and 1992/93. The bears taken by Cambridge Bay were mostly from northeastern Victoria Island. It was stipulated that the six bears allocated to Sachs Harbour would be for males and taken north of Norway Island (within the Northern Beaufort subpopulation).

In the negotiations for a management agreement for Viscount Melville Sound, the management area was adjusted and a quota of four was settled upon. Ulukhaktok was allocated a quota of four for Viscount Melville Sound in 1993-94. Beginning in the 1994-95 hunting season, a five-year moratorium on hunting polar bears in Viscount Melville Sound took effect because it was concluded that the population was overharvested. After that, a rotation took place between Cambridge Bay and Ulukhaktok, in alternate years, for a quota of four bears. Since Ulukhaktok had the last quota from Viscount Melville, the new rotation was scheduled to begin with Cambridge Bay in 1999-2000. Commencing in 2004/2005 the quota for Ulukhaktok and Cambridge Bay was set at four and three bears, respectively. That annual quota was thought to be less than the potential sustainable removal rate.

Harvest of the VM subpopulation has been below the quota for several years. Changing sea ice, distance needed to travel, challenging ice conditions and cost of travel to access bears are all cited as reasons (Larry Carpenter pers. comm. 2020). Changing sea ice conditions has made it difficult for Inuvialuit to rely on established IK for planning harvest activities (JS, 2015).

Protected Areas

There are currently no formal protected areas in the VM subpopulation but the area is very remote and limited human activities happen in the area. With decreased ice, shipping – especially that which includes ice breaking – is a potential conservation concern.

References

- Bethke, R., Taylor, M., Amstrup, S. and Messier, F. 1996. Population delineation of polar bears using satellite collar data. *Ecol. Appl.* 6:311-317.
- Brower, C.D., A. Carpenter, M.L. Branigan, W. Calvert, T. Evans, A.S. Fischbach, J.A. Nagy, S. Schliebe, and Stirling, I. 2002. The Polar Bear Management Agreement for the Southern Beaufort Sea: An evaluation of the first ten years of a unique conservation agreement. *Arctic.* 55:362–372.
- Canadian Wildlife Service (CWS). 2010. Northwest Territories – Inuvialuit Consultation Report on Polar Bear: Report on the Consultation in the Northwest Territories Inuvialuit Settlement Region in February 2009 on the Proposed Listing of Polar Bear as a Species of Special Concern Under the Federal Species at Risk Act. Canadian Wildlife Service, Yellowknife, NT. 121 pp.
- Carpenter, L. 2020. Personal communication at the Polar Bear Technical Committee meeting, February 3-6, 2020.
- Florko, R.N., G.W. Thiemann, and J.F. Bromaghin. 2020. Drivers and consequences of apex predator diet composition in the Canadian Beaufort Sea. *Oecologia* 194: 51-63.
- Florko, R.N., G.W. Thiemann, J.F. Bromaghin and E.S. Richardson. 2021. Diet composition and body condition of polar bears (*Ursus maritimus*) in relation to sea ice habitat in the Canadian High Arctic. *Polar Biology* 44:1445-1456.
- Joint Secretariat. 2015. Inuvialuit and Nanuq: A Polar Bear Traditional Knowledge Study. Joint Secretariat, Inuvialuit Settlement Region. xx + 304 pp. Available online: http://www.wmacns.ca/pdfs/394_polar-bear-tk-report-low-res.pdf.
- Joint Secretariat. 2017. Inuvialuit Settlement Region Polar Bear Joint Management Plan. Joint Secretariat, Inuvialuit Settlement Region. vii + 66 pp. https://www.nwtspciestatrisk.ca/sites/default/files/isr_polar_bear_joint_management_plan_2017_final.pdf
- Kingsley, M.C.S., Stirling, I., and Calvert, W. 1985. The distribution and abundance of seals in the Canadian High Arctic, 1980-82. *Canadian Journal of Fisheries and Aquatic Sciences* 42:1189–1210.
- Messier, F., Taylor, M. K. and Ramsay, M. A. 1992. Seasonal activity patterns of female polar bears (*Ursus maritimus*) in the Canadian Arctic as revealed by satellite telemetry. *J. Zool.* 226:219-229.
- Messier, F., Taylor, M. K. and Ramsay, M. A. 1994. Denning ecology of polar bears in the Canadian Arctic Archipelago. *J. Mammal.* 75:420-430.
- Schweinsburg, R. E., D. J. Furnell, and S. J. Miller. 1981. Abundance, distribution, and population structure of polar bears in the lower Central Arctic Islands. Wildlife Service Completion Report Number 2. Government of the Northwest Territories, Yellowknife, NT.
- Slavik, D., Wildlife Management Advisory Councils (Northwest Territories and North Slope), and Inuvialuit Game Council. 2009. Inuvialuit Knowledge of Nanuq: Community and Traditional Knowledge of Polar Bears in the Inuvialuit Settlement Region. Inuvialuit Joint Secretariat, Inuvik, NWT. 67 pp. Report and Unpublished Transcripts.
- Stern, H.L., and Laidre, K.L. 2016. Sea-ice indicators of polar bear habitat. *The Cryosphere* 10:2027–2041.
- Taylor, M. K., J. Laake, H. D. Cluff, M. Ramsay, and F. O. Messier. 2002. Managing the risk from hunting for the Viscount Melville Sound polar bear population. *Ursus* 13:185-202.
- Taylor, M. K., Akeagok, S., Andriashek, D., Barbour, W., Born, E. W., Calvert, W., Cluff, . D., Ferguson, S., Laake, J., Rosing-Asvid, A., Stirling, I., and Messier, F. 2001. Delineating Canadian and Greenland polar bear (*Ursus maritimus*) populations by cluster analysis of movements. *Can. J. Zool.* 79:690-709.

Western Hudson Bay (WH)

Status and Delineation

Hudson Bay is a relatively shallow inland sea that is ice covered in winter and ice free in summer (Hochheim *et al.* 2010). Although three subpopulations of polar bears (Foxye Basin, Southern Hudson Bay, and Western Hudson Bay) occur on the sea ice in winter and spring, they appear to be largely segregated during the open-water season (Derocher and Stirling 1990; Peacock *et al.* 2010; Viengkone *et al.* 2016). During the ice-free period, Western Hudson Bay polar bears exhibit strong fidelity to terrestrial summering areas in northeastern Manitoba (Stirling *et al.* 1977; Derocher and Stirling 1990; Cherry *et al.* 2013; Stapleton *et al.* 2014; Lunn *et al.* 2016). The current Western Hudson Bay subpopulation boundary is based largely on capture, recapture, and harvest of tagged animals (Stirling *et al.* 1977; Derocher and Stirling 1990, 1995a; Taylor and Lee 1995; Lunn *et al.* 1997).

The size of the Western Hudson Bay subpopulation was unknown until the 1990s (Derocher and Stirling 1995a). During the 1960s and 70s, the numbers of polar bears likely increased as a consequence of the closure of the fur trading post at York Factory, withdrawal of military personnel from Churchill, and the closure of hunting in Manitoba (Stirling *et al.* 1977; Derocher and Stirling 1995a). Derocher and Stirling (1995a) estimated the mean population size for 1978-1992 to be 1,000 (SE = 51). However, this estimate was considered conservative because the study had not covered the southern portion of the range east of the Nelson River (Calvert *et al.* 1995; PBSG 1995) and, therefore, for management purposes the population size was adjusted to 1,200 (Calvert *et al.* 1998). In 1994 and 1995, Lunn *et al.* (1997) expanded the capture program to sample animals to the Western Hudson Bay/Southern Hudson Bay management boundary and estimated abundance to be 1,233 (SE = 209) in 1995. Regehr *et al.* (2007) reported a decline in abundance from 1,194 (95% CI = 1,020-1,368) in 1987 to 935 (95% CI = 794-1,076) in 2004 and also documented that the survival rates of cubs, sub-adults, and old bears (>20 years) were negatively correlated with the date of sea ice breakup.

A mark-recapture distance sampling study resulted in an abundance estimate of 1,030 (95% CI = 754-1,406) in 2011 (Stapleton *et al.* 2014). During this survey, 711 bears were observed and more bears, particularly adult males, were observed in the coastal areas east of the Nelson River towards the Western Hudson Bay/Southern Hudson Bay boundary than were documented during the late 1990s (Stirling *et al.* 2004). Stapleton *et al.* (2014) suggested that a distributional shift may have negatively biased abundance estimates derived from capture samples. Mean litter size (cubs-of-the-year, 1.43 ± 0.08 ; yearlings, 1.22 ± 0.10) and number of cubs observed as a proportion of total observations (cubs-of-the-year, 0.07; yearlings, 0.03) were lower than those recorded for the neighboring subpopulations of Foxye Basin and Southern Hudson Bay, which is consistent with Western Hudson Bay having low reproductive productivity (Regehr *et al.* 2007; Peacock *et al.* 2010; Stapleton *et al.* 2014). The body mass of solitary adult female polar bears has declined over the past 42 years, which has likely contributed to declining reproductive success (Derocher and Stirling 1995b; Stirling *et al.* 1999; Sciuillo *et al.* 2016; Johnson *et al.* 2020; Lunn and McGeachy 2022).

Lunn *et al.* (2016) evaluated the demography and status of the Western Hudson Bay subpopulation for the period 1984-2011, using a Bayesian implementation of multistate capture-recapture models, coupled with a matrix-based demographic projection model, to integrate several types of data and to incorporate sampling uncertainty, and demographic and environmental stochasticity across the polar bear life cycle. Their analysis resulted in an estimate of 806 (95% CI = 653,984) for polar bears in the core area of study north of the Nelson River in 2011. Although both the aerial survey and capture-recapture estimates are broadly similar with overlapping confidence intervals, it is difficult to make direct comparisons because the geographical area covered differed. The aerial survey likely provides an accurate “snapshot” estimate of the total number and distribution of polar bears in the Western Hudson Bay management area at the time of the survey. The point estimate of abundance from the capture-recapture model represents the number of bears that move through the smaller, capture-recapture sampling area.

A second mark-recapture distance sampling study in 2016 was undertaken to update subpopulation status (Dyck *et al.* 2017). Pre-survey consultations with Nunavut Hunters’ and Trappers’ Organizations, Kivalliq communities, and with the Manitoba Department of Sustainable Development were conducted in order to use local and traditional knowledge in the study design. Dyck *et al.* (2017) reported the final estimate of abundance to be 842 bears (95% CI: 562–1121). Although not statistically significant from the previous aerial survey estimate, this difference represents an 18% decline in the point estimates of abundance from the 2011 and 2016 aerial surveys. Over the same period of time and using similar methods, Obbard *et al.* (2018) documented a 17% decline in abundance for the neighbouring Southern Hudson Bay subpopulation. Similar to observations from the 2011 survey, cubs-of-the-year and yearling cubs comprised a small proportion of the sample size (Dyck *et al.* 2017) and suggested that low reproductive performance of the Western Hudson Bay subpopulation has continued.

In late August and early September 2021, a collaborative aerial survey (Atkinson *et al.* 2022) was conducted following the survey designs, including study area stratification and transect spacing, used in the 2011 and 2016 WH surveys (Stapleton *et al.* 2014; Dyck *et al.* 2017). As was observed in these previous surveys, bears were concentrated along the coast and offshore islands, although both lone individuals and family groups were also regularly sighted inland, particularly within Wapusk National Park. Atkinson *et al.* (2022) reported an estimated abundance of 618 bears (95% CI=425-899). Comparison to aerial survey estimates from 2011 and 2016 suggests that WH may be decreasing in abundance; 40% and 27% reduction from 2011 and 2016, respectively. Further, post-stratification of the 2021 survey results by sex and age classes showed significant declines in the abundance of both adult female and subadult bears between 2011 and 2021. Atkinson *et al.* (2022) were unable to definitively conclude whether the finding of declining abundance in WH over the last decade, specifically that of adult females and subadults, was the result of reduced survival and recruitment, movement of bears into neighbouring subpopulations (emigration), harvest pressure, or a combination thereof. Based upon the multiple lines of evidence, they concluded that it was plausible that all these factors have contributed to some degree. Of particular concern, however, was their finding that the observed declines are consistent with long-standing predictions regarding the demographic effects of climate change on polar bears. Subsequent biopsy mark-recapture work conducted by ECCC and the OMNR has indicated that significant inter-annual movements occur across the WH and SH boundary along the Manitoba/Ontario coastline of western Hudson Bay (unpublished data). These initial data suggest that caution should be taken in interpreting aerial survey results as of seasonal and annual shifts in distributions of bears.

Harvest Management

The management of the Western Hudson Bay subpopulation is the shared responsibility of the Governments of Manitoba and Nunavut, Parks Canada Agency, the Nunavut Wildlife Management Board (NWMB), and the Wapusk Management Board. Whereas

there is no hunting of polar bears in Manitoba, in Nunavut, WH polar bears are harvested under a quota system for which the ultimate management responsibility lies with the Government of Nunavut (GN), as represented by the Minister of Environment. This responsibility is subject to the terms of the *Nunavut Agreement*, which established a system of ‘co-management’ for wildlife. Under the *Nunavut Agreement*, the Minister’s decision-making authority for wildlife management is shared with the NWMB and subject to strict requirements for consultation with Regional Wildlife Organizations and community-based Hunters and Trappers Organizations. This shared decision-making includes the establishment, modification, and removal of restrictions on the harvest of polar bears. Although any restrictions on the harvest of polar bears must be within limits set by the *Nunavut Agreement*, the NWMB or the Minister of Environment are permitted to restrict or limit Inuit harvesting but only to the extent necessary to affect a valid conservation purpose or to provide for public health or public safety. Due primarily to concerns for public safety, four increases in the total allowable harvest (TAH) of WH polar bears have occurred since 2011 that have raised the quota in Nunavut by 17 and the current total permissible removal to 42: Nunavut, 38; and, Manitoba, 4 (defense/accidental human-caused mortalities).

In September 2019, a new Nunavut Polar Bear Co-management Plan was approved by the Nunavut Wildlife Management Board, following five years of consultation and development with co-management partners in Nunavut. The Plan replaced the Memorandum of Understanding that had previously been in place. Concomitant with the approval of the new Plan, and in response to public and stakeholder feedback, Nunavut changed the sex-ratio of the harvest. Beginning with the 2019/2020 harvest season, up to 50% of a community’s quota can be harvested as females, which replaces the previous 2:1 male-biased harvest; no changes to existing community TAHs were made. Therefore, there is a potential that the biological risk of negative population outcomes due to harvest will increase because more adult female polar bears could be taken and these bears are the most important contributors to population growth. The GN is monitoring how this implemented change in harvest sex-ratio will affect the sex-ratio of harvested bears with the average female proportion of harvest in 2019/2020 until 2021/2022 equaling 28%. Females have been underharvested relative to the annual recommended quota by approximately 17.1% on average during 2017/2018 – 2021/2022 (5 yr) time period. In February 2022, the Polar Bear Flexible Quota System was replaced by the Harvest Administration and Credit Calculation System. The up to 1:1 harvest sex-ratio is still in place.

Protected Areas

Most of the known maternal denning area is protected within Wapusk National Park of Canada. Created in 1996, this 11,475 km² National Park is a remote wilderness area with no direct road access. Additional protection outside of the National Park is provided within the Churchill (8,500 km²) and Kaskatamagan (5,500 km²) Wildlife Management Areas designated under the Manitoba *Wildlife Act*. In February 2008, the polar bear in Manitoba was recognized as Threatened under the Manitoba *Endangered Species and Ecosystem Act*; which further ensures its protection, along with its habitat on both Crown and privately-owned land. The listing provides the ability to restrict development near critical habitat along the Hudson Bay coastline in Manitoba.

Indigenous Knowledge

Inuit report that the size of the Western Hudson Bay subpopulation has increased when compared to historical levels (McDonald *et al.* 1997; Tyrrell 2006; Nirlungayuk and Lee 2009).

From the 1930s through the 1960s, encounters with polar bears in the interior of the Kivalliq mainland and along the Kivalliq coast of Hudson Bay were rare (Nirlungayuk and Lee 2009; Tyrrell 2009). Within the last few decades, encounters with polar bears in the Kivalliq region have increased. Bears have also been observed near and within WH communities (Arviat, Chesterfield Inlet, Rankin Inlet, Whale Cove), resulting in more bear-human encounters and increased concerns for human safety and property damage. Based on both historical and recent observations, the general observation is that the Western Hudson Bay subpopulation has increased (Tyrrell 2006, 2009; Canadian Wildlife Service 2009; Nirlungayuk and Lee 2009; Henri *et al.* 2010; Kotierk 2012).

References

- Atkinson, S.N., Boulanger, J., Campbell, M., Trim, V., Ware, J., and Roberto-Charron, A. 2022. Aerial survey of the Western Hudson Bay polar bear subpopulation 2021. Final Report. Government of Nunavut, Department of Environment, Wildlife Research Section, Igloolik, Nunavut, Canada, 91 pp.
- Calvert, W., Taylor, M., Stirling, I., Kolenosky, G.B., Kearney, S., Crête, M., and Luttich, S. 1995. Polar bear management in Canada 1988-92. Pp. 61–79 In Wiig, Ø., Born, E.W., and Garner, G.W. (eds.). *Polar Bears: Proceedings of the Eleventh Working Meeting of the IUCN/SSC Polar Bear Specialist Group*. IUCN, Gland, Switzerland and Cambridge, UK.
- Calvert, W., Taylor, M., Stirling, I., Atkinson, S., Ramsay, M.A., Lunn, N.J., Obbard, M., Elliott, C., Lamontagne, G., and Schaefer, J. 1998. Research on polar bears in Canada 1993-1996. Pp. 69–91 In Derocher, A.E., Garner, G.W., Lunn, N.J., and Wiig, Ø. (eds.). *Polar Bears: Proceedings of the Twelfth Working Meeting of the IUCN/SSC Polar Bear Specialist Group*. IUCN, Gland, Switzerland and Cambridge, UK.
- Canadian Wildlife Service. 2009. Nunavut consultation report – Consultations on the proposed listing of the Polar Bear as Special Concern under the Species at Risk Act. Report submitted to the Nunavut Wildlife Management Board in accordance with Step 3.8 of the Memorandum of Understanding to Harmonize the Designation of Endangered Species under the Nunavut Land Claims Agreement and the Species at Risk Act, 249 pp. [available at: <http://assembly.nu.ca/library/Edocs/2009/001149-e.pdf>].
- Cherry, S.G., Derocher, A.E., Thiemann, G.W., and Lunn, N.J. 2013. Migration phenology and seasonal fidelity of an Arctic marine predator in relation to sea ice dynamics. *Journal of Animal Ecology* **82**:912–921.
- Derocher, A.E., and Stirling, I. 1990. Distribution of polar bears (*Ursus maritimus*) during the ice-free period in western Hudson Bay. *Canadian Journal of Zoology* **68**:1395–1403.
- Derocher, A.E., and Stirling, I. 1995a. Estimation of polar bear population size and survival in western Hudson Bay. *Journal of Wildlife Management* **59**:215–221.

- Derocher, A.E., and Stirling, I. 1995b. Temporal variation in reproduction and body mass of polar bears in western Hudson Bay. *Canadian Journal of Zoology* **73**:1657–1665.
- Dyck, M., Campbell, M., Lee, D.S., Boulanger, J., and Hedman, D. 2017. Aerial survey of the western Hudson Bay polar bear sub-population 2016. 2017 Final Report. Government of Nunavut, Department of Environment, Wildlife Research Section, Igloolik, Nunavut, Canada, 82 pp + 2 supplements.
- Henri, D., Gilchrist, H.G., and Peacock, E. 2010. Understanding and managing wildlife in Hudson Bay under a changing climate: some recent contributions from Inuit and Cree ecological knowledge. Pp. 267–289 In Ferguson, S.H., Loseto, L.L., and Mallory, M.L. (eds.). *A Little Less Arctic: Top Predators in the World's Largest Northern Inland Sea*. Springer, New York, New York, USA.
- Hochheim, K., Barber, D.G., and Lukovich, J.V. 2010. Changing sea ice conditions in Hudson Bay, 1980-2005. Pp. 39–51 In Ferguson, S.H., Loseto, L.L., and Mallory, M.L. (eds.). *A Little Less Arctic: Top Predators in the World's Largest Northern Inland Sea*. Springer, New York, New York, USA.
- Johnson, A.C., Reimer, J.R., Lunn, N.J., Stirling, I., McGeachy, D., and Derocher, A.E. 2020. Influence of sea ice dynamics on population energetics of Western Hudson Bay polar bears. *Conservation Physiology* **8**(1):coaa132, <https://doi.org/10.1093/conphys/coaa132>.
- Kotierk, M. 2012. Public and Inuit interests, Western Hudson Bay polar bears and wildlife management: results of a public opinion poll in western Hudson Bay communities. Department of Environment, Government of Nunavut, Iqaluit, Nunavut, Canada, 55 pp.
- Lunn, N.J., and McGeachy, D. 2022. Western Hudson Bay Polar Bear Program Annual Report – 2021/22. Unpublished Typescript, Environment and Climate Change Canada, Edmonton, Alberta, Canada, 46 pp.
- Lunn, N.J., Servanty, S., Regehr, E.V., Converse, S.J., Richardson, E., and Stirling, I. 2016. Demography of an apex predator at the edge of its range – impacts of changing sea ice on polar bears in Hudson Bay. *Ecological Applications* **26**:1302–1320.
- Lunn, N. J., Stirling, I., Andriashek, D., and Kolenosky, G.B. 1997. Re-estimating the size of the polar bear population in Western Hudson Bay. *Arctic* **50**:234–240.
- McDonald, M., Arragutainaq, L., and Novalinga, Z. (Compilers). 1997. *Voices from the Bay: Traditional Ecological Knowledge of Inuit and Cree in the Hudson Bay Bioregion*. Canadian Arctic Resources Committee, Ottawa, Ontario, Canada.
- Nirlungayuk, G., and Lee, D.S. 2009. A Nunavut Inuit perspective on Western Hudson Bay polar bear management and the consequences for conservation hunting. Pp. 135–142 In Freeman, M.M.R., and Foote, L. (eds.). *Inuit, Polar Bears and Sustainable Use: Local, National and International Perspectives*. Canadian Circumpolar Institute Press, Edmonton, Alberta, Canada.
- Obbard, M.E., Stapleton, S., Szor, G., Middel, K.R., Jutras, C., and Dyck, M. 2018. Re-assessing abundance of Southern Hudson Bay polar bears by aerial survey: effects of climate change at the southern edge of the range. *Arctic Science* **4**:634–655.
- PBSG [IUCN/SSC Polar Bear Specialist Group]. 1995. Summary of polar bear population status 1993. Pp. 19–24 In Ø. Wiig, Born, E.W., and Garner, G.W. (eds.). *Polar Bears: Proceedings of the Eleventh Working Meeting of the IUCN/SSC Polar Bear Specialist Group*. IUCN, Gland, Switzerland and Cambridge, UK.
- Peacock, E., Derocher, A.E., Lunn, N.J., and Obbard, M.E. 2010. Polar bear ecology and management in Hudson Bay in the face of climate change. Pp. 93–115 In Ferguson, S.H., Loseto, L.L., and Mallory, M.L. (eds.). *A Little Less Arctic: Top Predators in the World's Largest Northern Inland Sea*. Springer, New York, New York, USA.
- Regehr, E.V., Lunn, N.J., Amstrup, S.C., and Stirling, I. 2007. Effects of earlier sea ice breakup on survival and population size of polar bears in western Hudson Bay. *Journal of Wildlife Management* **71**:2673–2683.
- Sciullo, L., Thiemann, G.W., and Lunn, N.J. 2016. Comparative assessment of metrics for monitoring the body condition of polar bears in Western Hudson Bay. *Journal of Zoology* **300**:45–58.
- Stapleton, S., Atkinson, S., Hedman, D., and Garshelis, D. 2014. Revisiting Western Hudson Bay: Using aerial surveys to update polar bear abundance in a sentinel population. *Biological Conservation* **170**:38–47.
- Stirling, I., Jonkel, C., Smith, P., Robertson, R., and Cross, D. 1977. The ecology of the polar bear (*Ursus maritimus*) along the western coast of Hudson Bay. Canadian Wildlife Service Occasional Paper 33, 64 pp.
- Stirling, I., Lunn, N.J., and Iacozza, J. 1999. Long-term trends in the population ecology of polar bears in western Hudson Bay in relation to climatic change. *Arctic* **52**:294–306.
- Stirling, I., Lunn, N.J., Iacozza, J., Elliott, C., and Obbard, M. 2004. Polar bear distribution and abundance on the southwestern Hudson Bay coast during open water season, in relation to population trends and annual ice patterns. *Arctic* **57**:15–26.
- Taylor, M., and Lee, J. 1995. Distribution and abundance of Canadian polar bear populations: a management perspective. *Arctic* **48**:147–154.
- Tyrrell, M. 2006. More bears, less bears: Inuit and scientific perceptions of polar bear populations on the west coast of Hudson Bay. *Journal of Inuit Studies* **30**:191–208.

Tyrrell, M. 2009. West Hudson Bay polar bears: the Inuit perspective. Pp. 95–110 In Freeman, M.M.R., and Foote, L. (eds.). *Inuit, Polar Bears and Sustainable Use: Local, National and International Perspectives*. Canadian Circumpolar Institute Press, Edmonton, Alberta, Canada

Viengkone, M., Derocher, A.E., Richardson, E.S., Malenfant, R.M., Miller, J.M., Obbard, M.E., Dyck, M.G., Lunn, N.J., Sahanatien, V., and Davis, C.S. 2016. Assessing polar bear (*Ursus maritimus*) population structure in the Hudson Bay region using SNPs. *Ecology and Evolution* **6**:8474–8484.